Agenda Item 8.3

Projects and Activities Supported by ASCOBANS

Developing Guidelines for Cetaceanfriendly Marine Spatial Planning

Document 8.3 Draft Guidelines for Cetacean-sensitive Maritime Spatial Planning for the ASCOBANS Area

**Action Requested** 

Take note

Comment

Submitted by

Intersessional Working Group



# DRAFT GUIDELINES FOR CETACEAN-SENSITIVE MARITIME SPATIAL PLANNING FOR THE ASCOBANS AREA

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### **Explanatory Note:**

The draft high-level recommendations are intended to inform maritime spatial planning policy and practice at national and international levels. Their role is to provide orientation and guidance for Maritime Spatial Planning (MSP) practitioners and those sectors guided by MSP policy (e.g. renewable energy) who do not have technical expertise with regard to cetacean management and transboundary marine conservation. The review of threats to cetaceans and appropriate MSP responses is intended to provide an overview of the range of pressures facing cetaceans and possible MSP responses. The current text is based on an extensive review of the international scientific literature on cetacean conservation and maritime spatial planning, conducted as part of the preparation process. This draft (3.0) has benefitted from the comments and suggestions generously provided by expert reviewers in May / June 2023 and comments received at a technical workshop in June 2023.

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# 1. Introduction

# 1.1. Current Status and Policy Context

Maritime spatial planning (MSP) is an integrative policy instrument concerned with the coordination and management of human activities at sea, with the aim of facilitating the sustainable development of ocean resources and the protection of the marine environment (EU MSP Directive 2014, IOC & EC 2021). Ecosystem-based MSP is founded on the principle that all human activities at sea must be carried out in such a way that does not risk the integrity, resilience and health of marine ecosystems and is aligned with internationally agreed conservation objectives (EC & ECIEEA 2021). This document provides guidance on how to achieve cetacean-sensitive MSP; MSP that is aligned with the conservation and restoration of dolphin, whale and porpoise populations, in accordance with the Agreement on the Conservation of Small Cetaceans of the Baltic, North East Atlantic, Irish and North Seas (ASCOBANS 1992). The realisation of cetacean-sensitive MSP is dependent on a high degree of transboundary coordination and cooperation across countries, in the case of ASCOBANS including both EU and non-EU member states.

The ASCOBANS agreement came into force in 1994 against the background of concern regarding the conservation status of cetaceans, particularly in the Baltic and North Seas. Since 2008 the ASCOBANS area has been expanded to include the Irish Sea and parts of the Northeast Atlantic. ASCOBANS is an independent UN agreement under the framework of the Convention on Migratory Species (CMS 1979), administered by the United Nations Environment Programme (UNEP). The ASCOBANS range covers large parts of EU waters as well as all of UK, and extends into Norwegian and Russian waters. The corresponding CMS agreement for the Black Sea, Mediterranean Sea and contiguous areas of the Atlantic Ocean (ACCOBAMS) entered into force in 2001. The Parties to the ASCOBANS Agreement undertook "to cooperate closely in order to achieve and maintain a favourable conservation status for small cetaceans." (ASCOBANS 1992, Res. 2.1). Favourable Conservation Status (FCS) is defined under the Convention on the Conservation of Migratory Species of Wild Animals (CMS 1979, Art. 1) in accordance with four criteria:

- (1) population dynamics data indicate that the migratory species is maintaining itself on a long-term basis as a viable component of its ecosystems;
- (2) the range of the migratory species is neither currently being reduced, nor is likely to be reduced, on a long-term basis;
- (3) there is, and will be in the foreseeable future sufficient habitat to maintain the population of the migratory species on a long-term basis; and
- (4) the distribution and abundance of the migratory species approach historic coverage and levels to the extent that potentially suitable ecosystems exist and to the extent consistent with wise wildlife management;

FCS in relation to small cetaceans is further specified under ASCOBANS Resolution 8.5 (ASCOBANS 2020). Here it was agreed that: "the general aim should be to minimize (i.e. ultimately to reduce to zero) anthropogenic removals (i.e. mortality), and in the short term, to restore and/or maintain biological or management units to/at 80 per cent or more of the carrying capacity". All Parties to the ASCOBANS Agreement and all ASCOBANS range states are furthermore Parties to the Convention on Biological Diversity (CBD). Under the Kunming-Montreal Global Biodiversity Framework, agreed under the auspices of the CBD in December 2022 it was agreed that Parties would engage in integrated spatial planning focussed on addressing biodiversity loss and would: "ensure that all areas are under participatory, integrated and biodiversity inclusive spatial planning and/or effective management processes addressing land- and sea-use change, to bring the loss of areas of high biodiversity importance, including ecosystems of high ecological integrity, close to zero by 2030, while respecting the rights of indigenous peoples and local communities" (CBD 2022, Kunming-Montreal Global Biodiversity Framework, Target 1). Under the OSPAR Convention for the North-East Atlantic (including Greater North Sea and Celtic Seas), fifteen contracting Parties in

northern and western Europe are committed to the application of an ecosystem-based approach to marine management as well as to the development of tools that support an ecosystem approach, including MSP (OSPAR 2010a). Similarly, the nine Baltic Sea states are committed to the implementation of ecosystem-based MSP with the objective of reducing environmental pressures of sea-based human activities on the Baltic Sea ecosystem and strengthening the protection and restoration of marine species and habitats (HELCOM 2021, HELCOM-VASAB 2021). All ASCOBANS Parties are members of the International Whaling Commission (IWC) and have thereby committed themselves to the conservation of cetaceans, including for example the prevention of ship strikes, reduction of ocean noise and cetacean-sensitive development of renewables at sea, all of which are of direct relevance for MSP.

Cetaceans and their habitats are also 'highly protected' under European Union law<sup>1</sup>. The EU Habitats Directive commits Member States to ensuring a favourable conservation status for natural habitats and their wild flora and fauna (EC 1992). The EU Marine Strategy Framework Directive (MSFD, EU 2008) requires Member States to achieve and maintain a 'Good Environmental Status' for the marine environment. The EU Maritime Spatial Planning Directive (MSPD; Directive 2014/89/EU; EU 2014) requires the implementation of an ecosystem-based approach to the planning and management of European seas. The EU Biodiversity 2030 Strategy (EC 2020) commits Member States to achieving a target of at least 30% of the land and 30% of the sea under protection by 2030. Strict protection is envisaged for areas of high biodiversity value or potential, accounting for at least one third of all marine protected areas or 10% of EU sea area. These targets have since become a cornerstone of the Kunming-Montreal Global Biodiversity Framework. Under the ESPOO Convention adopted in 1997 (UNECE 2017) European countries are obliged to notify and consult each other on all major projects likely to have significant adverse environmental impacts across international boundaries. The provisions of this convention are very relevant in the marine context, where the adverse impacts of resource extraction and other economic activities must be considered from a transboundary, regional seas perspective (Pinarbasi et al 2020, Moodie & Sielker 2022). It is recognised, however, that European-level legal protections have to date not been sufficient to generate effective conservation, notably for cetaceans (Evans 2018, Carlén et al. 2021). Indeed, both the EU Green Deal (EC 2019) and the EU Offshore Renewable Energy Strategy (EC 2020) call for a massive upscaling of the offshore renewable energies, raising questions of compatibility with biodiversity and ecosystem conservation objectives.

Cetaceans in European waters face multiple threats arising from human activities. These include but are not limited to contaminants, habitat degradation, noise pollution, fishing bycatch and prey depletion. Threats vary in their severity both between species and across European sea-basins. For example, certain whale species are particularly vulnerable to collisions with ships in areas of high whale density such as the Bay of Biscay and waters of the Iberian Peninsula, whereas bycatch and habitat degradation due to eutrophication among other factors, are potential pressures for harbour porpoise in the Baltic Sea (ICES WGMME 2019). ASCOBANS covers all odontocetes (toothed whales) besides the sperm whale, all of which depend upon acoustic cues both to navigate and locate prey, whilst all but the harbour porpoise produce tonal sounds that are used in communication. Sound is their primary sense and thus they are particularly sensitive to underwater noise. In excess, these can lead to mass strandings, injury, disorientation, displacement and behavioural change (Nowacek et al. 2007, Mann & Teilmann 2013, Wisniewska et al. 2018, Bernaldo de Quiros et al. 2019). In extreme cases, such as the unmitigated removal and explosion of underwater ammunition (CMS 2016) and active sonar use, underwater noise can be fatal for cetaceans (Goertner 1982, Finneran and Jenkins 2012). At the same time, benthic ecosystems may benefit from restrictions on fishing activity (in particular scallop dredging and beam trawling) in, and around the vicinity of offshore windfarms (Coates et al. 2016, Roach et al. 2018) with potentially positive indirect impacts on cetacean populations. On the other hand, localised recovery of benthic ecosystems may coincide with increased fishing pressure elsewhere due to displacement effects, and uncertainty remains regarding the overall impacts of offshore wind on benthic communities (Dannheim et al. 2020).

<sup>&</sup>lt;sup>1</sup> EU currently applies to all ASCOBANS parties with the exception of the UK.

Policy responses need to be sensitive to the conservation requirements of individual species and be tailored to the specific conditions prevailing in each sea basin area. It should be recognised that the current intensification of economic activity at sea, is actively supported through the EU's "blue economy" agenda (European Commission 2023a), and ocean-basin specific policy decisions, such as the Ostend Declaration for the North Sea (Ostend Declaration 2023), and will place additional pressure on an already stressed marine environment. Pressures on cetaceans due to noise pollution associated with pile-driving during the construction phase of wind farms (Bailey et al. 2010, Branstetter et al. 2018, Benhemma-Le Gall et al. 2021, Huang et al. 2023), for example, may compound existing pressures arising from long-established maritime activities such as commercial fishing and shipping. Major pollution incidents and near misses (e.g. MV Fremantle Highway in August 2023) demonstrate the vulnerability of marine ecosystems and the potential for catastrophic impacts from single incidents.

Observed and predicted impacts of climate change on cetaceans include changes in distribution, range and migratory movements, increased inter-specific competition, changes in behaviour and reproduction, changes in the distribution of prey species, changes to marine ecosystems, and increased exposure to contaminants and toxic algae (Sousa et al., 2019; Evans and Waggitt, 2020; van Weelden et al., 2021; Kebke et al., 2022). Some of the predicted changes threaten the future survival of some cetacean populations (Tulloch et al., 2019), while observed changes have exacerbated the collapse of threatened species (Meyer-Gutbrod et al., 2021). The increasing frequency of marine heatwaves can cause an immediate and long-lasting decline in cetacean reproductive success through impacts on prey species and ecosystems (Gabriele et al., 2022). Climate change and marine heatwaves can have indirect impacts too, such as displacing cetaceans into areas where they are subjected to higher exposure to other negative impacts such as entanglement (Santora et al., 2020). While some species (e.g. harbour porpoise) may be able to adapt to climate change by moving to colder waters (van Weelden et al. 2021), others such as the white-beaked dolphin, may suffer significant habitat loss with potentially significant impacts for their conservation (Lambert et al., 2014; Nunny and Simmonds, 2020). Climate change projections indicate increased risks associated with extreme weather. Increased intensity of storm surges and coastal flooding pose risks to maritime infrastructure, such as windfarms, ships, pipelines and cables with potential severe impacts for cetaceans and their habitats. Risk management and disaster risk reduction will need to be integrated within MSP (European Commission 2023b). Changes in human behaviour due to climate change (such as changes to shipping routes in arctic seas) may also lead to greater impacts on some cetacean populations (Nunny and Simmonds, 2020). Due to their sensitivity to climate and ecosystem change, cetaceans can have an important role as 'sentinels'; indicators of the wider health status of marine ecosystems (Bossart et al. 2011, Wiliamson et al. 2021).

Under multiple stressor conditions exacerbated by climate change, critical thresholds are quickly reached and irreversible adverse impacts on cetacean populations may result (National Academies of Sciences, Engineering and Medicine 2017). The precise measurement of such thresholds remains a scientific challenge, due to the inherent complexity of marine ecosystems, the need for better interdisciplinary integration and the data limitations (Orr et al. 2022). Where such thresholds have been identified (e.g. with regard to underwater noise) it is imperative that they inform MSP. The accelerated deployment of offshore renewable energy risks adverse impacts on cetacean populations, unless comprehensive mitigation measures are implemented and adhered to (e.g., Gill 2005, Madsen et al. 2006). Close coordination between neighbouring countries is critical to manage cumulative impacts on cetaceans and their habitats (Maxwell et al. 2013, Halpern et al. 2015, Platteeuw et al. 2017). For this reason, the implementation of effective ecosystem-based maritime spatial planning is essential to ensure that human activities at sea are compatible with cetacean conservation objectives (e.g., Hammar et al. 2020, Carlucci et al. 2021). An ecosystem-based approach requires planners to work within the limits of marine ecosystems, whilst recognizing that humans are a component of that ecosystem (Curtin and Prellezo 2010). Planners should therefore have regard to the cumulative impact of the full range of human activities at sea rather than focusing solely on the additional impacts from emerging activities or individual planning proposals (see also Birdlife International 2022). Unfortunately, recent evaluations of maritime spatial plans for Baltic and North Sea states indicate that current practice falls short of delivering effective ecosystem-based

management and that more robust and thorough methodologies are required to assess likely cumulative effects and ecosystem sensitivity (Birdlife International 2022, WWF 2022).

Due to the complexity and dynamic nature of marine ecosystems and the interactions between marine ecosystems and human activities, it is in many cases not possible to scientifically pinpoint the critical threshold points at which significant adverse impacts may be expected to occur. Similarly, the cumulative impacts of multiple activities occurring simultaneously, or within a delimited geographical space are not easily quantified and decision-makers will need to work with a high degree of uncertainty. As a consequence of the above, it is imperative that the precautionary principle is adhered to, as enshrined in EU law<sup>2</sup> and committed to under both the OSPAR and HELCOM Conventions. This principle requires decision-makers to err on the side of caution, to take preventative action, even where full information is not available or certain risks are not quantifiable.

Effective conservation of small cetaceans requires a combination of area-based and threat-based measures (Evans 2018). Area-based measures focus on the protection of specific habitats and species within defined geographical boundaries. Such measures may take the form of marine protected areas (MPAs) and/or dedicated zoning, or Maritime Spatial Planning (MSP). Fundamentally, the ecological carrying capacity of individual areas and processes should guide such zonation. In some cases, networks of MPAs have been designated prior to the preparation of MSP. An ecologically coherent MPA network can, under these circumstances form the backbone for ecosystem-based MSP, around which human activities are planned, but only if they are applied with a focus upon areas and habitats that are important for a range of species. Improved monitoring and assessment are required to ensure the location and extent of protected areas are informed by current scientific data. Where MPA networks are not yet established, MSP may facilitate the identification of valuable and vulnerable marine habitats or ecosystems which may be subsequently designated as MPAs and/or given effective protected area status through MSP zoning measures. In order to effectively allow for ecosystem restoration, and provide sanctuary for stressed populations, protected areas should reduce and mitigate human activities with potential adverse impacts. In some cases, temporary or seasonal exclusions may be suitable whereas in others permanent exclusions are necessary, and again, MSP allows for the adoption of a more flexible approach for the management of ocean activities. Area-based conservation measures alone are, however, insufficient to ensure effective conservation. In addition, threat- or sector-based measures are required to address threats that are widespread and not specific to particular locations. Such threats include but are not limited to fisheries bycatch, ship strikes, contamination due to chemical pollution, underwater noise and marine litter. Threat-based measures are particularly necessary for highly mobile species, as is the case with cetaceans.

# 1.2. Ecosystem-based Maritime Spatial Planning and Cetacean Conservation

Maritime Spatial Planning (MSP) has evolved through a need to improve the management of the marine environment for the optimized use of marine resources, whilst simultaneously protecting the marine ecosystem (Ehler and Douver, 2009). Maritime spatial planning may be defined as a "public process of analyzing and allocating the spatial and temporal distribution of human activities in marine areas to achieve ecological, economic, and social objectives that usually have been specified through a political process" (Ehler & Douvere, 2009,18). The EU MSP Directive, similarly refers to MSP as "a process by which the relevant Member State's authorities analyse and organise human activities in marine areas to achieve ecological, economic and social objectives" (EU MSPD 2014, Art. 3). MSP essentially involves the establishment of a coordinated process that avoids, or minimizes, conflicts between different uses in the marine environment whilst maintaining ecosystem services that support these uses. MSP is concerned with both the regulation of uses of and claims on sea space and setting out medium and long-term integrated cross-sectoral visions for the future of marine space (Walsh et al. 2022).

<sup>&</sup>lt;sup>2</sup> The precautionary principle was introduced into EU law and policy via the 1992 Treaty on the European Union (the Maastricht Treaty) Article 130(2) (now Article 191 (2) of the Treaty on the Functioning of the European Union), which stipulated that the EU's environmental policy was to adhere to the precautionary principle.

MSP may vary according to the objectives under which the process is established. It is, however, possible to identify the following key characteristics:

- **Area-based**: MSP involves managing areas within which a variety of uses of the marine environment and resources occur.
- **Ecosystem-based**: the health, functioning and interaction of components of the marine ecosystem are fully considered in a systemic integrated manner.
- **Forward-looking**: the process is an anticipative one, not only looking at current activities but also future activities.
- Science-driven: all decisions are based on scientific information and evidence.
- **Transparent**: data and tools of different types support the decision-making process, and the information from these is freely available to stakeholders.
- **Participatory and integrated**: stakeholder participation and cross-sectoral integration is crucial in the entire process.
- Adaptive: activities are monitored, and plans are revised according to the observed effectiveness of actions and the receipt of new scientific information or evidence (see also Ehler 2021).

Although zoning remains an important element of MSP, multi-use is receiving increased attention in a number of national jurisdictions. Multi-use may be defined as "the joint use of resources in close geographic proximity by either a single user or multiple users" (Schupp et al. 2019). It is understood to represent a radical departure from the concept of exclusive resource rights to embrace the "inclusive sharing of resources and space by one or more users" (ibid.). Multi-use has the potential to reduce the demand for space and thus the impact of human activities on marine ecosystems (Depellegrin et al. 2019, Stancheva et al. 2022). One study has identified the potential for multi-use to optimally combine aquaculture and biodiversity objectives at the Italian Mediterranean coast (Venier et al 2021). The long-term vision of the Belgian maritime spatial plan calls for multiple-use to become the norm for all use of space within the Belgian North Sea by 2050 (Federal Public Service: Health, Food Chain Safety and Environment 2019). The Dutch Government has actively supported the development of multi-use pilot projects as part of a strategy aimed towards finding a balance between offshore wind, nature conservation and seafood production (Steins et al. 2021). Regulatory, technical and socio-economic factors, however, continue to present challenges to the widespread adoption of multi-use (also Stuiver et al. 2016). Whether multi-use can deliver improved conservation outcomes for cetaceans is uncertain.

Approximately half of the world's maritime spatial plans, to date, have been drafted in Europe (Friess and Grémaud-Colombier, 2021). As such, the European Union (EU) has taken a higher-level role, supported by framework legislation and high-level policy initiatives in seeking to ensure coherence and compatibility in MSP development by individual member states (European Commission, 2018). In addition to supporting the local establishment of MSP in various locations in Europe, the EU is actively involved in applying MSP to address transboundary cooperation (European Commission, 2018, Hassler et al 2018). Although MSP is, in essence, an area-based instrument, its role is not limited to zoning and area-based measures. It is increasingly recognised that MSP can play an important role in the coordination and sequencing of activities across time as well as space (as noted in the previous section). This is particularly relevant, for example, with regard to the coordinated sequencing of individual wind farm developments as well as with regard to seasonal restrictions for conservation purposes.

MSP can play an important role in providing a forum for dialogue and joint management of landbased pressures with an impact on the marine environment (Walsh 2021, Smith et al 2022). Close alignment between MSP, land-use planning and river basin management (e.g. under the EU Water

Framework Directive) is necessary to address, for example, issues of nutrient runoff from agriculture and domestic sources, i.e., integrated coastal zone and watershed management (e.g. Loiseau et al. 2012). Land-based pressures on the marine environment should be fully integrated within maritime spatial plans, in line with a 'One Space' integrated territorial planning approach (Kidd et al. 2019, ESPON 2020). MSP can also help to open a space for dialogue on fisheries management with a view to working with fisheries organisations to reduce bycatch and prey depletion. Marine ecosystem-based (EB) management (Long et al. 2015) and planning may be defined in different ways, leading at times to some confusion and a perception that it is a vague concept similar to sustainable development. An ecosystem-based approach, however, is legally well-defined<sup>3</sup> internationally and should be understood to imply that management practices and planning measures are informed by an understanding of ecosystem functioning as well as of the interactions between ecosystems and human activities. Ecosystem-based Maritime Spatial Planning (EB-MSP) is understood to require knowledge integration across scientific disciplines, governance integration across sectors and levels of government and transboundary integration across both the land-sea interface and political-administrative jurisdictional boundaries whether international or sub-national (Lieberknecht 2020, WWF 2021). EB-MSP furthermore requires dynamic and adaptive management (Duck 2012, Maxwell et al. 2015). It is recognised that marine ecosystems are inherently dynamic across multiple timescales, that it is necessary to respond to long-term trends such as climate change and biodiversity loss, as well as to seasonal variations in species distribution and other variables. Effective examples of such dynamic management approaches are shipping speed limits, or vessel avoidance of areas of whale presence, and other protective actions, put in place when seasonally migrating North Atlantic right whales are detected outside (spatially and temporally) of seasonal area closures (Van Parijs et al. 2009, Silber et al. 2012, Conn and Silber 2013).

MSP is an adaptive, cyclical process and, therefore, regulations and mitigation measures need to be revised regularly as new information becomes available. Monitoring is also required to ensure effectiveness, to highlight unexpected events or results, and to identify if, or when, specific mitigation measures or regulations become redundant, or need to be changed or enhanced. It is moreover becoming increasingly evident that MSP needs to become more responsive to external changes such as technological developments and shifts in political priorities due to geopolitical concerns and/or energy security issues - one major current example being the Ukraine conflict and marine activities in the Black Sea and Baltic Sea. At the same time, MSP should continue to be underpinned by a rigorous assessment of ecosystem impacts and a thorough evaluation of alternative scenarios. Key elements of an ecosystem-based approach to MSP were set out and adopted by the 72<sup>nd</sup> Meeting of the VASAB Committee on Spatial Planning and Development of the Baltic Sea Region (CSPD/BSR) and approved by the 50<sup>th</sup> Meeting of the HELCOM Heads of Delegation (HOD 50-2016) for the Baltic Sea area in 2016. The Guidelines for the implementation of an ecosystem-based approach in MSP in the Baltic Sea area present an important step toward a common understanding of how the ecosystem-based approach can be applied in drawing up a spatial plan for a sea area in accordance with spatial planning legislation in force in the Baltic Sea countries. Baltic Sea states have reaffirmed their commitment to the further development and implementation of EB-MSP in the Regional Baltic Maritime Spatial Planning Roadmap 2021-2030. In Table 1, these elements are set out in adapted form and their relevance for cetacean-sensitive MSP is elaborated. These elements of an ecosystem-based approach apply to all sea basins in the ASCOBANS area and should not be viewed as specific to the Baltic Sea.

<sup>&</sup>lt;sup>3</sup> HELCOM & VASAB 2016: <u>https://helcom.fi/helcom-at-work/groups/helcom-vasab-maritime-spatial-planning-working-group/</u>

Table 1: Ecosystem-based MSP Elements and their Relevance for Cetacean-Sensitive MSP (following HELCOM & VASAB 2016)

EB-MSP Element	Description	Relevance for Cetacean-Sensitive MSP
Best available Knowledge and Practice	The allocation and develop- ment of human uses shall be based on the latest state of knowledge of the ecosystems as such and the practice of safeguarding the components of the marine ecosystem in the best possible way.	This implies that planning measures and development projects should be informed by the best available rel- evant knowledge concerning cetacean populations and potential impacts and most effective cetacean conservation practice. This implies measures may need to be undertaken that incur additional costs for individual projects or types of development (e.g., use of adequate bubble curtains and similar technology to reduce noise pollution during wind turbine construction - see Amaral et al. 2020, Bellmann et al. 2020, Wursig et al. 2000).
Precaution	Far-sighted, anticipatory and preventive planning shall pro- mote sustainable use in ma- rine areas and shall exclude risks and hazards of human activities on the marine eco- system. Those activities that according to current scientific knowledge may lead to signifi- cant or irreversible impacts on the marine ecosystem and whose impacts may not be in total or in parts sufficiently pre- dictable at present require a specific careful survey and weighting of the risks.	The application of a precautionary approach means that activities with potentially significant adverse im- pacts on cetaceans require measures to prevent any potential impacts. Where the potential impacts of cer- tain activities are uncertain, regulators and planners should 'err on the side of caution', for example, give initial consent for a minimal, or lower volume of activity to begin with. Careful assessment and monitoring are required to ensure that such activities do not have an avoidable adverse impact on the conservation status of cetacean populations, prior to changing manage- ment regimes to increase levels of activity (with further monitoring of the effects of such increase).
Alternative	Reasonable alternatives shall	This principle requires the proactive assessment and
development scenarios	be developed to find solutions to avoid or minimise negative environmental and other im- pacts as well as impacts on ecosystem goods and ser- vices.	development of realistic alternatives to reduce adverse impacts. Alternative actions are frequently a component of Environmental Impact Assessments, but too often alternatives are not really realistic, feasible or may be deliberately worse than the original offering to avoid changes in planned activities (Steinemann 2001). An openness to alternatives requires a continuous questioning of the status quo. It is also imperative that development proposals are considered in relation to existing and other planned activities, rather than in isolation in order to allow for an assessment of cumulative impacts and potential alternatives (Burris & Canter 1997).
Identification of Ecosystem	In order to ensure a holistic evaluation of effects and po-	Potentially relevant ecosystem services may include the functional role of cetaceans in maintaining healthy
Services	tentials, ecosystem services need to be identified. Ecosys- tem services encompass all di- rect and indirect contributions of ecosystems to human well- being – or in short, the benefits people obtain from nature. They include provisioning, reg- ulating, cultural and supporting services <sup>4</sup> .	and resilient marine ecosystems as well socioeco- nomic values such as the value of cetaceans in terms of carbon capture (Pearson et al. 2022), as charis- matic species, boosting wildlife tourism (Parsons et al 2003, Pacheo et al. 2021) and helping to raise public awareness of marine conservation issues, or their in- trinsic value to the public (Scott & Parsons 2005, Naylor & Parsons 2018).

<sup>&</sup>lt;sup>4</sup>Millennium Ecosystem Assessment: <u>https://www.millenniumassessment.org/en/index.html</u>, IPBES: <u>https://www.ipbes.net/glossary-tag/ecosystem-service</u>

EB-MSP	Description	Relevance for Cetacean-Sensitive MSP
Element		
Mitigation	Measures are envisaged to prevent, reduce, and as fully as possible offset any significant adverse effects on the environ- ment of implementing the plan	Maritime spatial plans should follow a strict mitigation hierarchy (e.g. Arlidge et al 2018). If significant ad- verse impacts cannot be prevented, proactive ceta- cean-specific measures should be implemented to mitigate adverse impacts cetacean on cetacean pop- ulations and restoration measures Implemented to off- set any negative impacts, as mitigation measures rarely reduce a risk completely. These mitigation measures should be continuously monitored and eval- uated to ensure that they are effective, and modified if they are not fully reducing impacts to cetaceans.
Relational Un- derstanding	It is necessary to consider var- ious effects on the ecosystem caused by human activities and interactions between hu- man activities and the ecosys- tem, as well as among various human activities. This includes direct / indirect, cumulative, short / long-term, permanent / temporary and positive / negative effects, as well as interrelations including sea- land interaction.	Measures for cetacean conservation cannot be con- sidered in isolation but should always be viewed and assessed within the context of the wider social and ecological systems. In some cases, targeted cetacean conservation measures may impact adversely on other taxa or indeed other cetacean species. In order to make informed decisions, an understanding of these complex relationships is needed.
Participation and Communication	All relevant authorities and stakeholders as well as the wider public shall be involved in the planning process at an early stage. The results should be communicated to all stake- holders and made available to the wider public in a transpar- ent manner.	Inclusive and meaningful participation is required both to increase the acceptance and perceived legitimacy of plan measures, mitigate potential conflicts; and en- hance the knowledge base for decision-making. Con- sultation and communication should occur early in the MSP process and care should be taken to listen to all voices, not just the loudest. Conducting both quantita- tive and qualitative social science surveys to assess levels of support Is essential early in the process to identify possible areas of support or conflict (Bennett et al. 2017, Bennett 2019, Sanborn & Jung 2021).
Subsidiarity and Coherence	Maritime spatial planning with an ecosystem-based approach as an overarching principle shall be carried out at the most appropriate level of govern- ance and shall seek coherence between the different levels.	Plans at both subnational and national levels of gov- ernance are necessary to ensure both strategic over- sight and the necessary level of detail required to im- plement an ecosystem-based approach. Plans at each level should be coordinated in terms of their knowledge base and planned conservation and miti- gation measures.
Adaptation	The sustainable use of the ecosystem should apply an iterative process including mon- itoring, reviewing and evalua- tion of both the process and the outcome.	An adaptive, dynamic, iterative process is essential to ensure that continuous learning and reassessment take place and that planners and regulators can re- spond in a timely manner to new information (e.g. on cetacean population trends, impact severity or mitiga- tion options) and unanticipated external change fac- tors (e.g. major pollution incidents).

# 1.3. Building on Existing Good Practice

Cetacean-sensitive MSP is not well-established in Europe, or elsewhere, although there have been some significant efforts to devise assessment methodologies for the integration of MSP and cetacean conservation in the Ionian Sea (Mediterranean Sea basin) (Carlucci et al 2021). However, in the past two decades, the number of cases where the spatial allocation of parts of the ocean has been adapted specifically to the needs of cetaceans have been growing steadily, for example the adjustment of shipping lanes to avoid ship strikes within the framework of the International Maritime Organization (IMO) (Vanderlaan et al. 2008). The integration of marine conservation and maritime spatial planning, more broadly, is at an early stage. Nevertheless, there are examples of good practices which should inform the further development and implementation of cetacean-sensitive MSP. In particular, progress has been achieved in the development of standards and thresholds for underwater noise. The MSFD Common Implementation Strategy Technical Group on Underwater Noise has prepared recommendations for EU threshold values for both continuous and impulsive underwater noise (TG NOISE 2022a, b). Both sets of recommendations were endorsed by representatives of the EU and associated countries under the auspices of the Czech presidency of the EU in November 2022. The threshold values have been set to inform EU Member States in their determination of Good Environmental Status as per the EU MSFD. It is imperative that maritime spatial plans follow these recommendations in their regulation and spatial coordination of relevant activities. These recommendations are particularly relevant in relation to the construction, operation and decommissioning of offshore windfarms but also equally apply to established activities including shipping, dredging and mining, military operations, underwater acoustic research and seismic surveys during oil and gas exploration. For example, in the Netherlands, a detailed methodology for the assessment of the potential effects of windfarm construction on harbour porpoises has been developed under the umbrella of the Dutch Framework for Assessing Ecological and Cumulative Effects – KEC 4.0, 2021 (Heinis et al. 2022). Maritime spatial plans should ensure critical thresholds for underwater noise are not exceeded, based on best available scientific knowledge in best practice impact assessment methodologies. A wide range of technical options for complying with noise limits is available and indeed, technological advances are likely to lead to increasingly effective mitigation (OSPAR 2020, Koschinski & Lüdemann 2020). Mitigation measures should be continuously reviewed for efficacy and rigorous enforcement and monitoring are essential to ensure such standards are met in practice.

The ability to monitor cetacean movement and behaviour has increased significantly in recent decades but requires ongoing investment of resources on a large scale. Established monitoring methods include aerial surveys, individual photo-identification, passive acoustic mapping and satellite tagging. Effective mitigation of adverse impacts due to shipping and fishing activity (e.g., continuous underwater noise, collisions, bycatch), will require improved information on vessel movements. In particular, vessels below 12 m in length are not required to carry Vessel Monitoring Systems (VMS). As a result, mapping of recreational boat activity is largely lacking (Evans 2018). Small fishing vessels can have a large impact on bycatch in some regional sea areas (e.g. Baltic Sea: Morkunas et al. 2022, Iberian Peninsula: Pierce et al. 2020) and can cause injuries via ship strikes, and can contribute significantly to underwater noise impacts (e.g. Picciulin et al. 2022) especially when vessel speed is not regulated.

Maritime spatial planning should play an active role in a transition from static mapping of activities and marine mammals to dynamic ocean management based on near real-time high-resolution spatial data on both cetacean and vessel movements (Maxwell et al. 2015, Hazen et al. 2018). Examples of dynamic ocean management in practice include bycatch reduction initiatives (e.g., the New England scallop fishery, East Australian multispecies longline fishery and the Hawaii Turtlewatch programme). Rather than replacing existing zoning-based measures, dynamic management tools can complement existing measures. They are particularly suited to the conservation of highly-mobile species such as cetaceans. Their implementation, however, requires the development of necessary infrastructure and comprehensive monitoring systems (e.g., passive acoustic buoys, communication systems, aerial surveys, networks of experienced observers) and the development of agreed protocols to ensure consistent response by fishing and other vessels to alerts received, and strict enforcement systems to ensure compliance. Their suitability in areas of low cetacean density (e.g. the Baltic Sea) remains to be tested. A dynamic system which does not have a high level of coverage, monitoring and enforcement is little better than no protection at all. The incorporation of dynamic elements within existing MSP systems can, however, help to strengthen the link between the strategic level of MSP and the day-to-day management of marine activities and resources.

### 2. High-level Recommendations

#### General Principles for Cetacean-Sensitive MSP

- Maritime spatial plans should include measures to ensure a Favourable Conservation Status (ASCOBANS 1992) for cetaceans is maintained or achieved and ensure adverse impacts are mitigated following Best Available Techniques (BAT) and Best Environmental Practices (BEP) in order to minimize the overall impact. There should be an evaluation process to ensure that BATs and BEPs effectively achieve minimal impacts.
- II. Maritime spatial plans should be **aligned with the achievement of conservation objectives in accordance with existing commitments**, including the ASCOBANS Agreement, Sea Basin cetacean conservation plans (e.g. ASCOBANS 2009, 2016a) and, where applicable, EU legislation and/or Regional Seas Conventions. The EU Marine Strategy Framework Directive calls for the achievement and maintenance of Good Environmental Status for marine ecosystems to be accorded priority over other interests. Similarly, cetacean-sensitive MSP should prioritise the achievement and maintenance of Favourable Conservation Status for cetaceans.
- III. Cetacean-sensitive MSP have the following characteristics:
  - **Strategic direction**: MSP processes should be guided by an overall strategy or vision, that outlines how to work towards long-term goals aligned with conservation objectives.
  - **Spatial and temporal coordination**: Spatial planning is traditionally concerned with the spatial coordination of human activities. The dynamic nature of the marine environment requires that greater attention is paid to temporal coordination both seasonally and longer term. Both spatial and temporal coordination are essential components of MSP.
  - **Dynamic adaptation:** Maritime spatial plans should have the capacity to adapt to changes in the marine ecosystem as well as changes in our knowledge of such systems (e.g., in relation to changes in the distribution and mobility patterns of cetacean populations). They should be accompanied by thorough, independent monitoring, which, where feasible, is aligned with the MSFD and other relevant monitoring cycles. The policies and zoning provisions contained within maritime spatial plans should be subject to continuous and regular monitoring and revision at least every six years.
  - **Incremental planning:** Planning of human activities at sea should occur in increments to allow for assessment and evaluation, based on the latest monitoring data, on a stepby-step basis. The duration of increments is largely dependent on the activity and the knowledge status of the impact of such an activity. Where there are knowledge gaps and/or significant uncertainties, increments should be shorter.
  - Mitigation of adverse impacts: The mitigation hierarchy (avoid, minimise, remediate, offset) (Alridge et al. 2018) should be rigorously applied with respect to the projected impacts of all human activities occurring within the plan area. Offsetting should be used as a last resort and be nature positive, resulting in an overall benefit to the cetacean population affected which again should be established by thorough monitoring and evaluation (see Jacob et al. 2020).
  - Rigorous assessment of environmental impacts: Maritime spatial planning should be accompanied by a rigorous and thorough assessment of environmental impact at both plan (SEA) and project (EIA) levels prior to activities (see Wright et al 2013). Moreover, communication and coordination are required so that EIAs are not conducted in isolation and cumulative Impacts of multiple projects can be considered by managers.

Rigorous monitoring and evaluation after projects have been initiated, are required to assess the efficacy of mitigation and management methods, with feedback process to ensure that future SEAs, EIAs and MSP cycles are better informed.

- IV. Maritime spatial plans should be informed by a functional understanding of marine ecosystems including a recognition that all human activities should be planned and carried out in such a way that does not lead to adverse impacts on the marine environment and is compatible with achieving and maintaining healthy and biodiverse marine ecosystems. A functional understanding of marine ecosystems requires:
  - Identifying core ecosystem components and their interlinkages (species, habitats, processes).
  - Assessing the current conservation status of cetaceans and other taxa
  - Identifying recent and long-term trends in population change.
  - Assessing the **likely impacts of both existing and planned human activities** on the marine ecosystem.
  - Identify and assess the **risks posed by low-probability high-magnitude events** (e.g. major pollution incidents).
  - Identifying **critical knowledge gaps and degrees of uncertainty** in relation to both cetacean distributions and the impacts of pressures arising from human activities.
  - Estimating the **carrying capacity of the marine ecosystem** with respect to both individual activities and the cumulative impact of all current and planned human activities (Gusatu et al 2021).
  - Assessing the **compatibility of existing and planned human activities** with the conservation and restoration measures required to achieve **Favourable Conservation Status** for cetaceans and **Good Environmental Status** for the marine ecosystem.
- V. Maritime spatial plans should be informed by the **precautionary principle**. This implies that where adverse impacts are considered possible or likely (e.g. within the SEA report or equivalent), **zoning should be conditional only**, and subject to an assessment at project level, determining that significant adverse impacts are not likely to occur in this instance. Where scientific information is incomplete but adverse impacts are considered likely (based on available information), the activities in question should **not be granted consent, unless effective mitigation can be guaranteed**.
- VI. Maritime spatial plans should make explicit recommendations not only on where activities should and should not occur but also on when they should occur, taking account of seasonal variations in the spatial distributions and behaviours of cetaceans (e.g. Nachtsheim et al. 2020) and the cumulative impact of the co-occurrence of multiple activities (or instances of the same activity) occurring within a short period of time. Co-occurrence of impulsive noise events should be avoided wherever possible. Application of bubble curtains and other mitigation measures to reduce the absolute impulsive noise levels in line with established best practices is critical where impulsive noise cannot be avoided (e.g. Bundesministerium für Umwelt, Naturschutz und Reaktorsicherheit 2013<sup>5</sup>).

# Cetacean Conservation and Restoration

VII. Maritime spatial plans should make provision for an ecologically coherent network of extensive cetacean conservation areas. Their locations should be informed by an assessment of the spatial distribution and abundance of individual cetacean species, encompassing both breeding and feeding grounds (e.g. Gilles et al. 2009). The critical sites for all cetacean populations that have an unfavourable population status should be included in such zones. The conservation objectives should be designed in such a way as to improve the conservation status of the population concerned. Cetacean conservation areas may vary along a spectrum from restriction zones with regulations specific to one maritime

<sup>&</sup>lt;sup>5</sup> German Federal Ministry for the Environment, Nature Conservation and Nuclear Safety: Concept for the protection of harbour porpoises from noise impact during the construction of wind farms.

activity (e.g. speed limits for shipping) to strictly protected areas. **Close cross-sectoral coordination** with the relevant public authorities (e.g., ministries and/or environmental protection agencies) is necessary to ensure that **conservation areas are designated as marine protected areas (MPAs)**.

- VIII. Maritime spatial plans should engage not only in cetacean conservation but also in ecological restoration. Restoration may be defined as: "assisting the recovery of a degraded, damaged or destroyed ecosystem to reflect values regarded as inherent in the ecosystem and to provide goods and services that people value" (Martin 2017). Restoration is necessary where cetacean populations, habitats or prey populations have experienced long-term decline and/or acute short-term decline. Restoration may take active (e.g. species reintroduction, planting of seagrass meadows, saltmarsh restoration) or passive (e.g. setting aside large areas for natural regeneration) forms. Restoration can occur both inside and outside of protected areas and is not necessarily more successful in areas of low human impact (Fraschetti et al. 2021). The recently adopted Global Biodiversity Framework mandates "that by 2030 at least 30 per cent of areas of degraded terrestrial. inland water, and marine and coastal ecosystems are under effective restoration, in order to enhance biodiversity and ecosystem functions and services, ecological integrity and connectivity" (CBD 2022). At the EU level the EU Biodiversity Strategy (EC 2020) and Draft Nature Restoration Law (European Parliament 2023) require member states to restore at least 20% of their total marine (and terrestrial) territories, irrespective of the degradation status. All ASCOBANS Parties need to pay urgent attention to integrating these restoration goals into national MSPs.
  - IX. Protected areas should be included in maritime spatial plans, encompassing a differentiated zoning system including strictly-protected no-take zones with a minimum of human activity and complementary zones where a limited range of compatible activities are permitted. The boundaries between the zones should not necessarily be fixed. They can be dynamically managed in response to shifts in the distribution and health of relevant cetacean populations. In order to enact such dynamic management, MPAs require continuous independent monitoring and adaptive management based on the scientific results of this monitoring.
  - X. MSP zones designated for economic activities that could have a negative impact on cetaceans should not overlap with cetacean conservation areas and cetacean-relevant MPAs. Appropriate scientifically informed buffer zones (informed by the spatial impact of the respective economic activity) should surround the cetacean conservation and protected areas (Agardy et al. 2011). Several studies have shown that offshore wind turbine construction and seismic surveys can both have large-scale effects as underwater noise can carry across considerable distances and cause behavioural change as well as hearing damage (e.g. Kavanagh et al. 2019).
- XI. Maritime spatial plans should ensure **connectivity between critical breeding, resting and feeding sites**, as well as the wider network of relevant MPAs. It is imperative that wind farms, aquaculture, shipping routes and other human activities **do not act as barriers or impediments to cetacean movement** (Fontaine et al. 2007, Gusatu et al. 2021).
- XII. Where **adverse impacts are found to occur or** (in exceptional circumstances), unavoidable adverse impacts are expected to occur due to planned activities at certain locations, **remediation (direct compensation)** with a demonstrable overall positive impact on affected cetacean populations should be implemented. Subsequent monitoring is required to determine if the remediation actions have had a sufficient beneficial impact on cetacean populations that unavoidable adverse impacts are not only compensated for, but there is a **net benefit to the population**.

XIII. Where the **abundance or health of a cetacean population has declined over at least six years**<sup>6</sup>, the maritime spatial plan should detail how the actions within the plan will contribute to **reversing this trend** and **contribute to ecosystem restoration**. Such measures should be commenced within 12 months of plan adoption.

### Environmental Assessment

- XIV. Strategic Environmental Assessments (SEAs; EU 2001) used in MSP processes should explicitly include cetacean species, and map their habitats and connectivity corridors, in order to subsequently assess how the favourable conservation status is being impacted and ultimately to inform the maritime spatial plan. SEAs should further demonstrate alignment with internationally agreed conservation objectives.
- XV. **Project-level environmental impact assessments** should be conducted in a thorough and rigorous manner according to harmonised, scientifically informed, methodologies. Planning authorities should provide for **safeguards and oversight mechanisms** that effectively ensure that environmental impact assessments are conducted on an **objective basis**, **independent of commercial interests**.
- XVI. Maritime spatial plans should be accompanied by a **detailed spatially-explicit assessment of the cumulative effects** of human activities (both existing and planned) on cetaceans (e.g. Halpern et al. 2015, Halpern et al. 2018, Quemmerais-Amice et al. 2020). This assessment, to be published within the SEA report (or equivalent) should include the following:
  - **Sensitivity matrix** of likely anthropogenic pressures on individual cetacean species (e.g. windfarm noise impact during the construction period on harbour porpoises)
  - Identification of the degree and character of key threats at species level
  - **Computation/extrapolation of cumulative effect scores** for each pressure for each grid square within a **high-resolution spatial grid** (e.g. 1 x 1km or 500m x 500m) (Hammar et al. 2020).
- XVII. Cumulative effects assessments should be conducted for **multiple distinct planning** scenarios with differing intensities and spatial distributions of human activities. These scenarios should be plausible and, where possible and relevant, consider shifting policy priorities. The differences in the impacts of alternative planning scenarios should be made clearly visible. The preferred planning scenario should be selected to ensure minimal adverse impact.

# Information Sharing and Transboundary Cooperation

- XVIII. In order to ensure that maritime spatial plans and consenting procedures are informed by accurate and up-to-date information, it is imperative that **data and knowledge pertaining to the marine ecosystem and potential threats** are shared **among all stakeholders**. It is imperative that the data gathered in the course of project-level environmental impact assessments is **shared with MSP decision-making bodies**. Protocols and oversight mechanisms should be developed and implemented to **prevent the withholding of relevant information** that might influence the capacity of planners to make informed decisions. The **public interest in ensuring healthy and diverse ecosystems** should be placed above private concerns regarding commercially sensitive data.
- XIX. Maritime spatial plans have a **responsibility to educate users of marine space and other stakeholders**, so as to improve the **capacity for evidence-informed decision-making**. This means that information on the spatio-temporal distribution, abundance and population health status of cetaceans within the plan area should be provided within the plan itself (on maps and in text form). Maritime spatial plans should **provide clearly accessible**

<sup>&</sup>lt;sup>6</sup> This six-year time period is aligned with EU MSFD monitoring cycles

**information** on long-term trends in species abundance. Formal plans should be **accompanied by online maps with up-to-date information**.

- XX. Maritime spatial plans should take **explicit account of transboundary impacts**. The current status of the **cetacean species and regional populations** (e.g. North Sea, Belt Seas and Baltic Proper harbour porpoises) should be considered, rather than solely the spatial distribution and abundance of cetacean species within the plan area (e.g., EEZ and/or coastal waters). In line with the Espoo Convention, maritime spatial plans should consider the **impact of current and planned activities in neighbouring jurisdictions.**
- XXI. Maritime spatial plans should include **commitments to coordinated planning and monitoring efforts**. **Monitoring methodologies** should be harmonised across the ASCOBANS Area. A **regional seas approach**<sup>7</sup> is recommended to ensure transboundary coordination and coherence of planning, environmental assessment and monitoring efforts.
- XXII. Where individual maritime spatial plans only cover parts of the national waters, such as the coastal zone or the Exclusive Economic Zone, a **consistent and coherent approach** should be adopted. The **same categories for cetacean conservation areas should be used at the various national and sub-national levels** and the cetacean management approaches in each plan should be integrated and based on a **common evidence base**.
- XXIII. The terms of reference of the **ASCOBANS Working Group on MSP** should be extended to encompass a coordination role in the development of **common assessment and monitoring methodologies for cetacean-sensitive MSP** and the sharing of relevant cetacean conservation expertise. The Working Group should **liaise and collaborate**, where possible and practical, with the WGs of other relevant IGOs, such as HELCOM, ICES, OSPAR.

### 3. Overview of Assessment of Cetaceans Impacts from Selected Sectoral Activities

In the following, we provide an overview assessment of three of the most significant maritime activities that can be regulated via MSP: offshore renewable energy, vessel traffic (shipping and boating) and fisheries including aquaculture. In each case, the specific threats, impacts on cetaceans and potential policy responses are outlined. Where relevant, we also address potential positive impacts of the above activities on cetaceans. Maritime spatial planning must consider a wider range of activities and claims on sea space including military use, sand and gravel extraction, oil and gas exploration and research activity. Further details on the broad spectrum of specific threats and pressures are found in Chapter 4.

# 3.1 Offshore Renewable Energy

The planned expansion of offshore renewable energy (ORE) capacity is projected to have significant impacts on marine ecosystems. Selecting appropriate locations for offshore wind farms (OWFs) and other forms of ORE requires careful evidence-informed ecosystem-based planning. It is imperative that site selection is informed and guided by ecological criteria. To date, the spatial distribution of installed and planned ORE has been determined by national policy and regulatory factors as much as by considerations of suitability and carrying capacity. It is anticipated that technological developments (e.g. floating turbines and increasing deployment of tidal, wave, and other forms of energy generation) will impact on the overall spatial distribution of ORE across northern and western Europe, potentially counterbalancing the existing concentration of activity in the North Sea and southwest Baltic Sea. ORE impacts and potential impacts on cetaceans are summarised in Table 2 below.

<sup>&</sup>lt;sup>7</sup> <u>https://www.unep.org/explore-topics/oceans-seas/what-we-do/regional-seas-programme</u>

Threats	Impacts	Policy measures
Impulsive noise im- pacts during OWF construction (see 4.4.1 below)	Disturbance leading to behav- ioural change, displacement (e.g. Benhemma-Le Gall et al. 2021).	Fine-scale spatial and temporal coordination to prevent co-occurrence of impulsive noise events. Use of BAT and BEP mitigation techniques such as double bubble curtains to reduce noise impacts. Application and rollout of alternative floating turbine foundations to avoid pile driving.
Continuous noise impacts of maintenance vessels (see 4.2.5 below)	Disturbance leading to, behav- ioural change, displacement (e.g. Stöber & Thomsen 2021)	Independent monitoring and continuous assessment. Restrictions on vessel and trip numbers, and vessel speeds, including seasonal restrictions as appropriate. Regulation and management of how service vessels moor offshore, ensuring that vessels minimize noise emission at all times.
Physical barrier effects due to offshore wind farms and wave devices	Potential impacts on habitat con- nectivity and cetacean mobility in areas of high ORE density (e.g. central North Sea) (Gussatu et al. 2021)	Rigorous assessment of cumulative effects at the sea-basin scale to ensure that barrier effects do not occur
Collision risk from tidal turbines	Overlap between high energy sites and important foraging ar- eas for cetaceans, leading to physical injury or death; also dis- placement from important feed- ing habitat (Benjamins et al. 2015)	Independent monitoring and continuous assessment. Systems for temporary shut- downs when animals come too close.

Table 2: Summary of adverse impacts of ORE on cetaceans

Table 3: Summary of potentia	I benefits of ORE for cetaceans
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Potential benefits for	Effects	Policy Measures
cetaceans		
Habitat enrichment due to	Positive impacts on benthic	Imposition of strict restrictions on
restrictions on fishing activity in	habitats and communities.	fishing activity in and adjacent to
OWFs.	Potential indirect impacts on	OWFs. Continuous monitoring to
	cetacean populations (Coates et	ascertain direct and indirect
	al. 2016, Roach et al. 2018).	ecosystem impacts
Offshore structures, including	Positive impact on cetacean	Decommissioned wind turbines
wind turbine foundations can act	populations (e.g. Fernandez-	may be left in situ to reduce noise
as artificial reefs and become	Betelu et al. 2022)	impact and allow for continued
important foraging grounds for	*	use as artificial reefs
cetaceans		

# 3.2 Vessel traffic (Shipping and Boating)

Shipping has significant adverse impacts on marine ecosystems and cetacean populations in particular. The volume of shipping traffic in the ASCOBANS area is high and projected to increase in coming decades, notably fast service vessels for the offshore wind industry which emit particularly high levels of underwater noise. Especially high densities of shipping traffic are found in the southern North Sea and English Channel. Unfortunately, maritime spatial plans and related policies have, to date, failed to effectively regulate the volume of shipping activity, irrespective of concerns regarding the carrying capacity and health of marine ecosystems. Enhanced transboundary cooperation and regulation may be necessary to ensure that critical thresholds are not exceeded. Shipping activity is a major source of continuous underwater noise, with long-term implications for the health of cetaceans at both individual and population levels. Shipping poses a significant risk due to the potential for large quantities of contaminants to enter the marine ecosystem as a result of major pollution incidents. Maritime spatial plans should take into consideration that increased volumes of

shipping will lead to increased probability of major pollution incidents and other accidents occurring. Shipping activity furthermore poses a significant collision risk for cetacean species. Whereas detailed monitoring data are available through AIS for large commercial vessels, it is increasingly evident that motorised recreational boats have significant adverse impacts, including disturbance effects, collision risk and as a contributor to continuous underwater noise.

Threats	Impacts	Policy measures
Continuous underwa-	Disorientation, feeding disrup-	Independent monitoring and enforcement of
ter noise (see 4.2.5	tion and reduced energy intake,	noise thresholds; speed restrictions, rerouting
below)	behavioural change, displace-	of shipping lanes and establishment of buffer
	ment (e.g. Findlay et al. 2023)	zones between sensitive areas and shipping
		lanes; incentives for quiet ship and propeller
		design
Collision risk (see	Risk of lethal and sub-lethal in-	Rerouting of shipping lanes; speed restrictions,
4.2.6 below)	jury, behavioural change, dis-	dynamic management measures (where
	placement (e.g. Peltier et al.	applicable)
	2019, Ritter & Panigada 2019)	
Disturbance from	Disorientation, behavioural	Guidance and regulation for wildlife tourism
recreational boating	change, displacement, lethal	operators and recreational boat users
and wildlife tourism	and sub-lethal injury (e.g. Peel	
(see 4.2.7 below)	et al. 2018, Olaya-Ponzone et	
	al. 2023)	
Risks of oil /	Potential acute and long-term	Rigorous assessment of cumulative effects at
chemical / hazardous	impacts on cetacean	the sea-basin scale to ensure such risks are
substance pollution	population health from pollution	minimised; speed restrictions.
from container ship	incidents (e.g. Wan et al. 2022)	
accidents (see 4.1.1		
below)		

Table 4: Summary of adverse impacts of shipping and boating on cetaceans

# 3.3 Fisheries and Aquaculture

Both fisheries and aquaculture pose significant risks for cetaceans in European waters. Bycatch, the unintended capture of non-target fish species and marine mammals in fishing nets represents the biggest threat to cetaceans globally (Elliot et al. 2023). Commercial fishing also contributes to the depletion of prey species (Pierce et al 2022). Depleted fish stocks have a direct impact on the viability of cetacean populations and, together with other factors, limit the scope for cetacean populations to recover and achieve a favourable conservation status. Fishing activity, furthermore, is a major contributor to underwater noise and disturbance of the seabed. Aquaculture poses significant risks due to contaminant pollution, eutrophication and entanglement in anti-predator nets (HELCOM 2018, Mazzariol et al. 2018, Carballeira Brana et al. 2021).

Threats	Impacts	Policy measures			
Bycatch (see	Cetacean mortality and injury	Seasonal restrictions, independent monitoring,			
4.1.2 below)		restrictions on the use of certain types of gear			
Prey depletion	Impact on health of cetacean	Quotas and restrictions on critical fish species,			
(see 4.1.3	populations	restrictions on the use of certain types of gear			
below)					
Underwater	Disorientation, behavioural	Speed restrictions, restrictions on the use of certain			
noise	change, displacement	types of gear, independent monitoring			
Introduction of	Acute and long-term impacts	Site-specific management and monitoring, site			
contaminants	on cetacean population	selection following strict ecological criteria			
(see 4.1.1	health from aquaculture				
below)					

Table 5: Summary of adverse impacts of fisheries and aquaculture on cetaceans

### 4. Threats to Cetaceans and Appropriate MSP Measures

As outlined above, cetaceans face a wide range of threats from a broad spectrum of human activities. There is considerable variation in terms of the relative significance of individual threats both geographically, across the ASCOBANS range area, and between cetacean species. Table 6 below, provides an overview of the key threats to the conservation status in each of the regional sea areas within the ASCOBANS range. Threats are classified as High (red), Medium (amber) or Low (green). This overview assessment is based on more detailed work carried out by the ICES Working Group on Marine Mammal Ecology (ICES WGMME 2019). Only those threats classified as medium or high for at least one regional sea area are included in this overview. In the final two right-hand columns of Table 3, each threat is classified by sector as well as by spatial distribution. The range of applicable MSP measures is very different for those threats that might be considered to be widespread (found across a sea-basin with limited spatial differentiation), relatively location specific (associated with activities at certain locations or found within certain geographical areas but which cannot be easily localised to specific point sources) and highly location specific (associated with specific point source activities). In the text below, each threat is described in detail with reference to relevant scientific literature. Potential MSP measures and opportunities for cross-sectoral coordination are highlighted in each case.

Threats	Regional Sea				Cetacean Species <sup>8</sup>	Sector	Spatial Distribution
	Baltic Sea <sup>9</sup>	Greater North Sea	Celtic Seas	Bay of Biscay & Iberian Peninsula			
Contaminants	Н	H	Н	н	harbour porpoise, killer whale, bottlenose dolphin, common dolphin, striped dolphin, long-finned pilot whale	Land- sea	Widespread
Habitat degradation	М	L	L	L	harbour porpoise	Land- sea	Relatively location- specific
Litter (including plastics and discarded fishing gear)	L	M	Μ	М	harbour porpoise, dolphins, minke whale, beaked whales, bottlenose whale, Risso's dolphin	Fishing	Widespread
Sonar	Н	M	Н	Η	harbour porpoise, minke whale, long- finned pilot whale, killer whale, bottlenose whale, beaked whales	Military	Relatively location- specific
Seismic surveys	H	M	H	Η	harbour porpoise, minke whale, bottlenose whale, beaked whales, long- finned pilot whale, killer whale, bottlenose dolphins, common dolphin, white-beaked dolphin, Atlantic white- sided dolphin, Risso's dolphin	Oil and gas	Relatively location- specific

Table 0. Overview matrix of theats, spatial distribution, species and sectors	Table 6: Overview ma	trix of threats, spatia	al distribution, s	pecies and sectors
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<sup>&</sup>lt;sup>8</sup> Note that only cetacean species covered by ASCOBANS are included in this table. Larger cetaceans such as sperm whales and fin whales are also impacted by some of the threats detailed in Table 6.

<sup>&</sup>lt;sup>9</sup> Including Belt Seas and Kattegat.

Threats	Regio	nal Sea			Cetacean Species <sup>8</sup>	Sector	Spatial Distribution
	Baltic Sea <sup>9</sup>	Greater North Sea	Celtic Seas	Bay of Biscay & Iberian Peninsula			
Pile-driving	М	М	М	0	harbour porpoise, minke whale, bottlenose dolphin	Offshore wind	Highly location- specific
Underwater Explosions	Н	М	0	0	harbour porpoise, minke whale, bottlenose dolphin	Military / offshore wind	Highly location- specific
Shipping (noise)	М	Μ	М	Μ	harbour porpoise, minke whale, bottlenose dolphin,	Shippin g	Relatively location- specific
Collision with ships	L	М	М	н	Minke whale, bottlenose dolphin	Shippin g	Relatively location- specific
Collision with tidal turbines	L	L	Μ	L	Harbour porpoise, bottlenose dolphin, common dolphin, Risso's dolphin, minke whale	Offshore renewab le energy	Highly location specific
Overfishing of prey species	М	М	М	Μ	harbour porpoise, minke whale, bottlenose dolphin, common dolphin, white-beaked dolphin, Atlantic white-sided dolphin	Fisherie s	Widespread
Removal of non-target species (by- catch)	Η	Η	Η	Н	harbour porpoise, minke whale, common dolphin, white-beaked dolphin, Atlantic white- sided dolphin, Risso's dolphin, striped dolphin, bottlenose dolphins	Fisherie s	Widespread
Disturbance (e.g. wildlife watching and recreational boating)	L	Μ	M	М	coastal bottlenose dolphin	tourism	Relatively location- specific

# 4.1. Widespread Threats

# 4.1.1. Contaminants

**Spatial Distribution:** high across all regional seas, highest in the Baltic and North Seas **Species most impacted:** harbour porpoises, bottlenose dolphins, killer whales **Sources of threat:** marine pollution incidents, wastewater treatment plants, coal-fired power stations, leaching from buildings and household materials, agriculture, urban runoff, atmospheric deposition of industrial plant emissions, military Installations and equipment, maritime sources (e.g., shipping, oil spills), PCBs (e.g. HELCOM 2018, OSPAR 2022).

Multiple contaminants, present in the marine environment, are known to have negative impacts on cetaceans. The contaminants of highest conservation concern are polychlorinated biphenyls (PCBs) which are known to have a negative effect on cetacean reproduction and health. Although production has been banned since 2001, they persist in marine ecosystems and food chains. Within the ASCOBANS area, the highest levels of PCBs in the marine environment occur in the Baltic Sea

(HELCOM 2010, 2018), and Greater North Sea, followed by the Celtic Seas and Bay of Biscay (OSPAR 2010, 2017). The cetacean species with the highest levels of PCB contamination in Europe include killer whale, bottlenose dolphin, striped dolphin and harbour porpoise (Jepson et al. 2016). Heavy metals are also known to cause a range of negative effects on marine mammals including neurological damage, organ damage, reduced reproductive success, and death (Das et al. 2002). Although heavy metals are in decline in many European seas, they remain a significant problem in some areas including the North Sea and Baltic Sea (EEA 2019). The average life expectancy of Baltic and North Sea harbour porpoises has been found to be dramatically reduced due to exposure to contaminants, severely impacting on the ability of the populations to reproduce (Sonne et al. 2020, Siebert et al. 2007).

### Appropriate MSP and policy responses:

Recognition that the wildlife health and fitness status of cetacean populations is poor as a result of high contaminant load and that therefore the overall resilience and ability to adapt to stressors is much reduced. Cumulative impact assessments should guide MSP. Need for stricter regulation of wastewater treatment plants and other sources of pollution and water quality monitoring under WFD and MSFD. This issue should be highlighted in maritime spatial plans but needs to be tackled also elsewhere.

# 4.1.2. Bycatch

**Spatial Distribution:** high across all regional seas, critical in the Baltic Sea **Species most impacted:** harbour porpoise, common dolphin, minke whale, and humpback whale **Sources of threat:** commercial fisheries

Bycatch remains the single largest threat to cetaceans globally, with an estimate of at least 300,000 cetaceans killed each year (Elliot 2020) and can impact all cetacean species. In the ASCOBANS area the species of greatest concern are harbour porpoise, common dolphin, minke whale, and humpback whale. Bycatch data may be insufficient to estimate bycatch rates for many areas and fisheries due to insufficient reporting compliance, monitoring and data collection (Hanke et al. 2020, Murphy et al. 2021, Dolman et al. 2021, Pierce et al. 2021).

# Harbour Porpoise

Bycatch of this species is of primary concern in the ASCOBANS area, with the main gear types involved being gill nets (Lusseau et al. 2023). Annual porpoise bycatch in the UK was estimated at 1098 animals for 2017 (Northridge et al. 2018), while bycatch in the Celtic Sea ecoregion (Subarea 7 only) represented between 1.1 and 2.4% of the porpoise population estimate for that area (ICES, 2018) and was potentially above the 1% precautionary environmental limit recommended by ASCOBANS as an indication that bycatch levels may have an impact on the population (ASCOBANS 2016b). Estimated bycatch mortality rates for the Iberian Peninsula also appear to suggest that the number of porpoises killed annually is likely to be unsustainably high (Pierce et al. 2021).

# Common Dolphin

Bycatch of common dolphins has been reported in pelagic trawl and purse seine fisheries, 'very high vertical opening' bottom-pair trawl fisheries, as well as (bottom-set) gillnets and long-lines (Murphy et al., 2021). Annual bycatch mortality levels across the NE Atlantic have been estimated in the hundreds or low thousands from independent observer programmes, although not all fisheries have been assessed (ICES 2016a, Murphy et al. 2013). Since 2005, overall numbers of common dolphin strandings have been increasing along the coasts of Ireland, the UK, and France (Murphy et al. 2021) and in 2016 ICES advised the European Commission that bycatches of common dolphins may be unsustainable (ICES 2016a). This was followed by a call for emergency measures to reduce bycatch in the Bay of Biscay, and recommendations for mitigation measures from ICES (2020, 2021, 2022).

### Minke Whale

Bycatch of minke whales in the ASCOBANS area primarily focuses on entanglement in static fishing gear particularly static pot fisheries (Northridge et al. 2010). Up to 30 minke whales are likely to be entangled in fishing creel lines in Scotland each year, with an estimated fatal entanglement rate of 2.3% per annum for the west coast of Scotland (Leaper et al. 2022).

### Humpback Whale

Bycatch of humpback whales in the ASCOBANS area also focuses on entanglement in static pot fisheries, with 6 humpback whales estimated entangled in fishing creel lines in Scotland each year (Leaper et al. 2022). Estimated mortality rates of humpback whales in fishing gear in Scotland were judged not to be sustainable and Scotlish inshore waters may act as a high mortality sink for NE Atlantic humpback whales (Ryan et al. 2016). Humpback whales within the ASCOBANS area may also be subject to entanglement risks in other areas of their migratory range. A scar analysis of 379 humpback whales in Iceland revealed that at least 24.8% of individuals had a history of prior entanglement when first encountered (Basran et al. 2019).

### Appropriate MSP and policy responses:

Concerted efforts should be made to improve the monitoring of fisheries and bycatch. Dynamic management measures should be piloted and implemented within maritime spatial plans. Continuous monitoring focussing on medium to high-risk fisheries is required. Mitigation can be achieved via targets set in consultation with fishermen. Temporary closures of areas of high bycatch should be considered (Dolman et al. 2016, Evans 2019).

# 4.1.3. Overfishing of Prey Species

Spatial Distribution: medium threat across all regional seas

**Species most impacted:** harbour porpoise, minke whale, bottlenose dolphin, common dolphin, whitebeaked dolphin, Atlantic white-sided dolphin **Sources of threat:** commercial fisheries

While a direct impact of overfishing of prey species on cetaceans is often difficult to prove due to most cetacean species having varied diets, the collapse of herring in the North Sea during the 1960s and the overfishing of sand eel around the Shetland Islands in the 1990s were both implicated in the concurrent decline of harbour porpoise in those regions (Reijnders 1992, Borges & Evans, 1997). Indeed, the collapse of the sand eel stock in the northern North Sea may well explain a large southward shift in harbour porpoise distribution in the North Sea between 1994 and 2005 (Hammond et al. 2013). Harbour porpoises are particularly vulnerable to prey depletion due to their high metabolic demands (Wisniewska et al. 2016, Booth 2020). Other cetacean species have also been shown to be vulnerable to prey depletion whether from overfishing or other causes (Bearzi et al. 2008, Ward et al. 2009, Ford et al. 2010, Kershaw et al. 2021, Cunen et al. 2021). Long-term declines have been recorded for many commercial fish species in the ASCOBANS area. Declines in sand eel, cod and whiting have been recorded in the Greater North Sea, sole and cod in the Irish Sea, and whiting in the Celtic Seas (OSPAR 2010, 2017, ICES 2022). Cod in the Kattegat has declined markedly in the past 50 years and herring in the Gulf of Riga (Baltic Sea Region) and sprat in the western Baltic (2001-2016) have been assessed as having unfavourable status (HELCOM 2018). Herring stocks in the Celtic Sea and West of Scotland/Ireland have also severely declined since the 1960s (Marine Institute 2021). The overall impact of these, and other, stock declines of important prey species for cetaceans in the ASCOBANS area are still poorly understood.

### Appropriate MSP and policy responses:

- Fisheries regulations (ICES, CFP, European Commission 2023c), MPAs with internationally agreed no-take zones
- This issue should be highlighted in marine spatial plans but needs to be tackled also via sectoral measures.
- Management of recreational fishing.

# 4.1.4. Litter (incl. plastics and discarded fishing gear)

**Spatial Distribution:** Medium threat across all regional seas with the exception of the Baltic Sea (low threat)

**Species most impacted:** various deep diving whale species (e.g. sperm whale, beaked whales) **Sources of threat:** lost/discarded fishing gear, land-based sources

Marine plastic litter is a major pollutant of marine habitats and is now ubiquitous in the ASCOBANS Area (OSPAR 2010, 2014, 2017, HELCOM 2018). The major sources of marine plastic litter are from land-based sources and from the fishing industry. Impacts of macro litter on cetaceans focuses on the ingestion of macro plastics by deep diving species such as sperm whales (*Physeter macrocephalus*), beaked whales and some oceanic dolphins (Baulch & Perry 2014, Lusher et al. 2018, Fossi et al. 2018), and in at least some cases the ingested plastic has resulted in the death of the animal (Fossi et al. 2018). Entanglement in fishing gear is an issue for some baleen whale species in the ASCOBANS Area such as minke whale (*Balaenoptera acutorostrata*) and humpback whale (*Megaptera novaeangliae*). Such entanglements may involve either active fishing gear or lost/discarded gear (netting, ropes).

### Appropriate MSP and policy responses:

Fisheries regulations, regulations targeting the producers of goods that end up as marine litter in order to aim for a circular economy (EU 2019), public awareness, beach and estuary clean-ups. This issue should be highlighted in maritime spatial plans but needs to be tackled via sectoral measures.

### 4.2. Relatively Location-Specific

#### 4.2.1. Habitat degradation

Spatial Distribution: medium in Baltic Sea, otherwise low

#### Species most impacted: harbour porpoise

**Sources of threat:** disturbance of the seafloor, through sand and gravel extraction, dredging, bottom trawling, and eutrophication leading to habitat depletion. The likely primary cause in the Baltic Sea is eutrophication (nutrient pollution and resulting anoxic areas and harmful algal blooms) due to run-off from agriculture, wastewater treatment and other land-based sources.

The development of offshore structures for oil and gas extraction, aquaculture and renewable energy may result in physical loss of habitat for cetaceans through removing feeding habitat and/or excluding cetaceans from preferred feeding areas either on a temporary or long-term basis, an alternative view is that offshore structures have the potential to act as new habitat for prey species and as de facto protected areas for prey species and cetaceans. Coastal habitat loss may result from ports and harbour developments and infilling or enclosure to create new land or develop tidal renewable energy (Waters and Aggidis 2016). Bottom trawling and scallop dredging can irreversibly alter or remove some types of benthic habitats such as oyster reefs (Pogoda 2019) and has had the widest impact on seabed habitats of any human activity (Eigaard et al. 2017).

Species-specific habitat loss may result from climate change and especially so for cold temperate and arctic species (MacLeod 2009, Chambault et al. 2018). Exclusion from preferred habitat on a short- or long-term basis can result from disturbance due to vessel traffic, water sports, underwater noise (e.g. ship noise, ADDs on aquaculture sites, seismic surveys, or offshore renewable energy developments) (Campana et al. 2015, Götz, and Janik 2013, Findlay et al. 2018, Kavanagh et al. 2019), or from chemical pollution events (Fisher et al. 2016). The pollution of coastal and transitional waters by phosphates and nitrates from land-based agricultural activities is a problem in the ASCOBANS area, with the Baltic Sea being of particular concern (EEA 2019). The consequences of coastal water enrichment (combined with the effects of warming seas) include algal blooms, which may impact coastal ecosystems and prey species, as well as potentially directly impacting cetaceans through through release of algal toxins. Toxic algal blooms have been known to cause mass mortality in cetaceans in the United States and elsewhere (Fire et al. 2015, Haüssermann et al. 2017).

Cetacean habitats can be degraded by a wide range of activities, while some impacts are local, others are widespread. Benthic habitats can be extensively degraded or permanently altered by scallop dredging and bottom trawling (Eigaard et al. 2017, Rijnsdorp et al. 2018), and dredging for sand and aggregates (Teaca et al. 2019, HELCOM 2018). Habitat degradation may reduce the value of habitat for cetaceans, or in extreme cases result in effective habitat loss. The degradation or loss of benthic habitat can result in the direct loss of foraging opportunities but can also have impacts on the wider marine ecosystems on which cetaceans depend. Pelagic habitats can also be degraded by noise pollution, chemical contaminants, litter, vessel traffic, and overfishing.

### Appropriate MSP and policy responses:

Restrictions on agricultural runoff, improved water quality monitoring under WFD (European Court of Auditors 2016) and protection of marine and coastal waters under EU MSFD. Large-scale nature restoration of relevant habitats, including those for prey species (e.g. reefs, seagrass). This issue should be highlighted in maritime spatial plans (land-sea interactions).

### 4.2.2. Sonar

**Spatial Distribution:** high in Baltic, Celtic Seas, Bay of Biscay and Iberian Peninsula **Species most impacted:** beaked whales (Cuvier's beaked whale, Sowerby's beaked whale, etc) **Sources of threat:** military use for navigation, detection of submarines, use in research, offshore wind turbine siting, and by the oil and gas industry.

The impact of mid-frequency (1-10 kHz) active sonar (primarily used by the military to detect submarines) on cetaceans has been identified as a significant conservation threat for deep-diving cetaceans (Evans and Miller 2004, Parsons 2017). In addition to disturbance and physical injury, the use of military mid-frequency active sonar has been linked to mass mortality in beaked whales and some other cetacean species linked to behavioural and physiological responses to sound (Bernaldo de Quirós et al. 2019, Parsons 2017, Jepson et al. 2013). Such mass mortality events can have potentially significant conservation impacts on local populations (Dolman et al. 2010, Brownlow 2018).

# Appropriate MSP and policy responses:

- Restrictions on sonar use in areas of cetacean abundance or vulnerability
- Independent monitoring
- Close exchange and coordination with military (e.g. nationally and through NATO).

# 4.2.3. Seismic surveys

**Spatial Distribution:** high in Baltic, Celtic Seas, Bay of Biscay and Iberian Peninsula **Species most impacted:** harbour porpoises, baleen whales and deep-diving whales **Sources of threat:** oil and gas exploration, offshore wind siting

Seismic surveys are the primary survey method currently in use by the oil and gas industry for locating deposits beneath the sea floor. They are also used in site surveys prior to the installation of rigs, wind farms and tidal energy installations. Seismic airguns produce loud impulse sounds with source levels up to 260–262 dB re 1  $\mu$ Pa-m (Thomsen 2009). Most of the energy produced by air gun arrays is in the low-frequency range below 300 Hz (Richardson et al. 1995), but some noise at higher frequencies is also generated.

Baleen whales and deep diving whales are thought to be most impacted by seismic surveys, with disturbance and disruption to foraging noted at tens of kilometres from the source (Nowacek et al. 2007, Miller et al. 2009), and a risk of permanent or temporary auditory injury to animals at close range (Nowacek et al. 2007, Southall et al. 2019). Strandings of deep-diving cetaceans have been linked to seismic survey activity (Castellote and Llorens 2015; McGeady et al. 2016), although direct evidence linking the two is elusive. Deep diving whales, and in particular beaked whales, may be particularly susceptible to the impacts of seismic noise due to their specific habitat requirements,

restricted habitat availability and the extreme physiological limits at which they live (Tyack et al. 2006, Barlow and Gisiner 2006, Wright et al. 2011).

Reaction to seismic surveys has also been noted in dolphins and porpoises (Thompson et al. 2013) and temporary auditory injury has been recorded in porpoises (Kastelein et al. 2017). Seismic surveys may last for weeks or months, and a single seismic survey is capable of ensonifying thousands of square kilometres of ocean, leading to potential masking of communication signals over wide areas (Clark and Gagnon 2006, Sutton et al. 2013). The effects of such chronic noise exposure on cetaceans over ocean basin scales is poorly understood but alteration to vocalisation rates and sightings rates have been reported (Clark and Gagnon, 2006; Kavanagh et al., 2019).

**Appropriate MSP and policy responses:** Avoidance and mitigation measures (see Parsons et al. 2019, Wright and Cosentino 2015), environmental impact assessments to ensure minimal impact, independent monitoring, and time period-specific to prevent co-occurrence with other impulsive noise events are recommended. Alternatives to seismic surveys include vibroseis (BOEM 2014).

### 4.2.4. Shipping Noise

**Spatial Distribution:** Medium across all regional seas (the busiest shipping routes in the ASCOBANS Area are the southernmost part of the North Sea and the English Channel, the southwestern Baltic, and across the outer part of the Bay of Biscay)

Species most impacted: harbour porpoise, whales (particularly baleen whales)

### Sources of threat: commercial shipping

Shipping is the most widespread source of noise pollution in our ocean (OSPAR 2009) and is recognised as a chronic, habitat-level stressor (Williams et al. 2020). Large ships typically have sound source levels of 160-220dB re 1µPa @ 1m at frequencies of 2-100Hz (Richardson et al. 1995, NRC 2003). As with all chronic noise, the impacts of shipping noise on cetaceans are poorly understood but impacts may include elevated stress levels (Rolland et al. 2012), exclusion from preferred habitat (Carome et al. 2022), masking of communication sounds, and masking of natural ocean sounds (Clark et al. 2009, Weilgart 2017, Erbe et al. 2019). Shipping noise primarily impacts low-frequency species such as baleen whales, but dolphins and porpoises can show strong avoidance reactions to ships (Dyndo et al. 2015) and smaller vessels (Pirotta et al. 2015). The busiest shipping routes in the ASCOBANS Area are the southernmost part of the North Sea and the English Channel, the southwestern Baltic, and across the outer part of the Bay of Biscay (OSPAR, 2010, 2017, Evans et al. 2011, HELCOM 2018).

### Appropriate MSP and policy responses:

Speed reduction and rerouting (IMO 2023). Rerouting of shipping lanes to reduce noise impact has been successfully conducted by Canadian authorities (via the IMO) in the Bay of Fundy at the southeastern Canadian coast (Vanderlaan et al. 2008). Dynamic management measures should be piloted and implemented within maritime spatial plans. Adoption of IMO ship design guidelines, and national guidelines where available (e.g. those of the German 'Blue Angel' quality mark for environmentally-friendly products<sup>10</sup>), in accordance with best available technology (BAT) principle.

# 4.2.5. Collision with ships

**Spatial Distribution:** high in the Bay of Biscay, medium in the Greater North Sea and Celtic Seas **Species most impacted:** baleen whales, sperm whale<sup>11</sup> **Sources of threat:** commercial shipping

Collisions between ships and cetaceans have mostly been reported in large whales (e.g. fin whale and sperm whale) and involving fast-moving ships (>10 knots) (Laist et al. 2001, Vanderlaan and

<sup>&</sup>lt;sup>10</sup> https://produktinfo.blauer-engel.de/uploads/attachment/de/Flyer\_BE\_Seeschiffe\_web.pdf

<sup>&</sup>lt;sup>11</sup> Note these species are not covered by the ASCOBANS Agreement.

Taggart 2007). Areas of highest collision risk are where high densities of large whales overlap with areas of high shipping density. Within the ASCOBANS area, the Bay of Biscay is the area of highest whale collision risk (Evans et al. 2011). Ferries, sailing yachts, passenger vessels and whale-watching boats accounted for most of the identified whale strikes in the International Whaling Commission (IWC) Ship Strike database (Winkler et al. 2019). High-speed ferries have been of particular concern in cetacean collision injuries and deaths (Ritter et al. 2019). Small vessels can also cause physical injury or death to cetaceans, including dolphins and porpoises (Feingold & Evans 2014, Peel et al. 2018, Olaya-Ponzone et al. 2023).

### Appropriate MSP and policy responses:

Speed reduction and rerouting (IMO (2023). Dynamic management measures should be piloted and implemented within maritime spatial plans.

# 4.2.6. Disturbance from Recreational and Wildlife Tourism Activity

**Spatial Distribution:** medium threat in Greater North Sea, Celtic Seas, Bay of Biscay and Iberian Peninsula

Species most impacted: bottlenose dolphin, harbour porpoise

Sources of threat: recreational activity and wildlife tourism

The increasing popularity of water sports and coastal activities, including whale and dolphin watching (Constantine and Bejder 2007, OSPAR 2010, 2017), is likely to lead to increased disturbance to cetaceans. At its most extreme, collisions between water sports and nature-watching vessels and cetaceans can result in the death or injury of individuals (Feingold & Evans 2014, Peel et al. 2018, Olaya-Ponzone et al. 2023). Other impacts include disturbance to animals during foraging, resting, socialising or breeding (Schaffar et al. 2009, Marino et al. 2012, Pirotta et al. 2015, Hoarau et al. 2020, Currie et al. 2021) from vessel presence or noise pollution (Burnham et al. 2021). Such disturbance may lead to reduced energy intake and increased energy expenditure with possible impacts on reproductive success and survival (Pérez-Jorge et al. 2016, Senigaglia et al. 2016). Coastal cetacean species are most likely to be affected by marine tourism activities. There is some evidence that cetaceans can adapt to the presence of whale-watching vessels, but this may be context-, or individual-specific (New et al. 2013, Di Clemente et al. 2018). The degree to which populations are affected by disturbance depends on the nature of the population, with closed populations most sensitive, while large, open populations with no food limitations are able to better tolerate disturbance (New et al. 2020). Disturbance may also alter the balance in marine ecosystems, leading to lower biodiversity and fewer 'rare' species (Sutton et al. 2021).

### Appropriate MSP and policy responses:

Regulations and international codes of practice have been established with relevant operators (e.g. International Whaling Commission Whale Watching Handbook<sup>12</sup>). Seasonal speed restrictions have been introduced within some protected areas.

# 4.3. Highly Location-Specific Threats

# 4.3.1. Pile-driving / wind farm construction

**Spatial Distribution:** medium threat in North Sea, Baltic Sea, Kattegat and Belt Seas, and Celtic Sea (e.g. Gusatu et al. 2021), projected to increase significantly in coming years. **Species most impacted:** harbour porpoise **Sources of threat:** wind farm construction

Pile-driving is used to insert metal pilings into the seabed to support offshore installations such as wind turbines. It can produce loud impulse sounds with source levels up to 243-257 dB re 1  $\mu$ Pa-m (Thomsen 2009). Most of the energy produced by pile driving is in the low-frequency range below

<sup>&</sup>lt;sup>12</sup> <u>https://wwhandbook.iwc.int/en/</u>

500Hz, but some noise at higher frequencies is also generated (Tougaard et al. 2009a; Dähne et al. 2013). Harbour porpoises have been observed to show negative reactions to pile driving at distances of 20 km or more (Tougaard et al. 2009a, Benhemma-Le Gall et al. 2021). Although porpoises have been noted to return after the construction phase of such developments (Brandt et al. 2018), during construction, porpoises and other cetaceans may be effectively excluded from preferred foraging or breeding areas. A planned increase in the construction of offshore wind and tidal energy sites within the ASCOBANS area may lead to impacts on porpoises and other cetaceans from pile driving noise, with cumulative impacts possible from adjacent developments. A considerable increase in potential impacts from planned offshore wind farm expansion in the North Sea is predicted by 2050 (Guşatu et al. 2021).

### Appropriate MSP and policy responses:

Project-level environmental impact assessments with independent oversight are essential, along with a need for careful temporal and spatial planning at the regional /national level to minimise the co-occurrence of impulsive noise events. There is a need for continuous independent monitoring to minimise adverse impacts on cetacean populations and to ensure cetaceans return following the construction phase. Need for an incremental adaptive planning approach. In accordance with Best Available Techniques and Best Environmental Practices, alternative non-percussive pile-sinking methods and floating turbine technology should be applied where possible. Double bubble curtains can serve as noise dampening mitigation measures. It is critical that established noise mitigation thresholds are being maintained as offshore wind turbines grown increasingly larger.

### 4.3.2. Underwater Explosions

**Spatial Distribution:** High threat in the Baltic Sea, medium threat in the Greater North Sea, Belt Seas and Kattegat

Species most impacted: harbour porpoise, minke whale, bottlenose dolphin

**Sources of threat:** site clearance prior to construction of marine installations (also due to controlled explosion of WWII ordnance)

Explosion of military ordinance has significant potential impacts on harbour porpoises. Blasting may also be required for site clearance prior to the construction of marine installations. Explosive blasts are one of the strongest point sources of any man-made sound. Source levels vary with the type and amounts of explosives used, and the water depth at which the explosion occurs. Source levels range from 272 to 287 dB re 1  $\mu$ Pa zero to peak at 1 m distance (for 1–100 lb. TNT) (Thomsen 2009). Low frequencies are generated with most energy between 6–21Hz (Richardson et al. 1995). Of extra concern is that blasting also produces shockwaves capable of killing or severely injuring marine mammals in the vicinity of the blast (Ketten 1995).

### Appropriate MSP and policy responses:

Project-level environmental impact assessments with independent oversight are essential, along with a need for careful temporal and spatial planning at the regional/national level to minimise the co-occurrence of impulsive noise events. Mitigation measures include acoustic deterrent devices to scare away animals prior to explosions and double bubble curtains to dampen noise levels (Koschinski & Lüdemann 2020) Continuous independent monitoring is also required to minimise adverse impacts on cetacean populations and to ensure cetaceans return following noise exposure. There is a need for an incremental adaptive planning approach.

# 4.3.3 Collisions with Tidal Turbines

**Spatial Distribution:** Medium threat in the Celtic Seas and northern North Sea, otherwise low **Species Most likely to be impacted**: Harbour porpoise, bottlenose dolphin, common dolphin, Risso's dolphin, minke whale

**Sources of Threat:** turbines for tidal energy generation

Concerns have been expressed over potential risk to seals and cetaceans of collisions with tidal turbines (Benjamins et al., 2015; Sparling et al., 2015; Onoufriou et al., 2019). Areas where turbines

exploit tidal energy have been deployed already include the Orkney Islands (Greater North Sea), Strangford Lough in Northern Ireland and off the coast of West Wales (Celtic Seas) with many further areas proposed, particularly around the UK. Potential interactions have been anticipated especially for harbour porpoise, which commonly utilise tidal stream environments, as well as some dolphins and seals (Benjamins et al., 2016; Waggitt et al., 2017). In those areas of the European Atlantic where tidal currents are strong, there are good prospects for deployment of tidal turbines as a source of offshore renewable energy. The deployment of tidal turbines is still very much in the experimental demonstration phase. So far, no actual incidents have been reported, with tagged seals taking aversive action (e.g. Sparling et al., 2018; Joy et al., 2018). With emphasis on improving energy security and alternative renewable energy sources, there are prospects for greatly increased deployment of tidal turbines.

### 5. Future Outlook: Towards Cetacean-Sensitive Maritime Spatial Planning

It is not unusual for maritime spatial plans to include long-term visions, providing a future perspective on the use of marine space in thirty- or forty-years' time. Indeed, maritime spatial plans should be accompanied by clear statements of strategy, outlining how the individual regulatory steps contained within the plan are envisaged to work together with sectoral measures to achieve desired outcomes. Based on current trends, a continued increase in maritime economic activity across the ASCOBANS area with a corresponding increase in risk to cetacean populations, is likely. Existing pressures will be compounded by the impacts of climate change, with significant impacts on the health and integrity of marine ecosystems, including cetaceans and other taxa. Increased volume and density of human activities and resulting pressures will increase the risk of high-magnitude low frequency events with severe and acute impacts on cetacean populations. Based on current legislation and policy initiatives, there is potential for significant progress in the implementation of ecosystem-based maritime spatial planning with a much closer alignment between maritime spatial planning, ecosystem management and marine protected area designation and management. Protected areas alone, however, will not be sufficient to achieve cetacean conservation objectives. Increased regulation and reduction in the intensity of sectoral uses is necessary to ensure to ensure the medium- and long-term viability of cetacean populations.

The use of maritime space will be subject to ongoing evaluation based on ecological criteria, including systematic assessments of cumulative impacts and analysis of short-, medium- and long-term trends. Technological advances will likely lead to more effective mitigation measures and increased opportunities for the multiple-use of marine space. The realization of cetacean-sensitive MSP will require a firm commitment by Parties and their relevant national agencies, as well as early leaders in the private sector (e.g. shipping, offshore wind) and other critical stakeholders. Investment in adequate capacities (e.g. monitoring), assessment methodologies and transboundary coordination at both strategic and operational levels will be essential to coherent, coordinated evidence-based approach. Cetacean-sensitive MSP will require a combination of both active and passive area-based conservation and restoration measures, seasonal restrictions, mitigation and remediation to allow cetacean populations to achieve and maintain a favourable conservation status.

The further development and implementation of cetacean-sensitive MSP represents a test case of scientifically-informed ecosystem-based MSP. Achieving favourable conservation status for cetaceans should be viewed in terms of their role as a key indicator of the health and resilience of marine ecosystems more generally, and as an essential step to achieving international biodiversity targets.

### List of Abbreviations

**ACCOBAMS:** Agreement on the Conservation of Cetaceans of the Black Sea, Mediterranean Sea and contiguous Atlantic area

**ASCOBANS**: Agreement on the Conservation of Small Cetaceans of the Baltic, North East Atlantic, Irish and North Seas

BAT: Best Available Technology

**BEP:** Best Environmental Practice

**CBD:** Convention on Biological Diversity

CINEA: European Climate, Infrastructure and Environment Executive Agency

CMS: Convention on the Conservation of Migratory Species of Wild Animals

**CSPD/BSR:** Committee on Spatial Planning and Development of the Baltic Sea Region (of VASAB)

dB re 1 µPa: technical measurement of sound pressure level

**EB-MSP:** Ecosystem-Based Maritime Spatial Planning

**EB:** Ecosystem-Based

**EC:** European Commission

**EEA:** European Environment Agency

EIA: Environmental Impact Assessment

EU CFP: European Union Common Fisheries Policy

EU MSFD: European Union Marine Strategy Framework Directive

**EU MSPD:** European Union Maritime Spatial Planning Directive

EU WFD: European Union Water Framework Directive

EU: European Union

HELCOM: Baltic Marine Environment Protection Commission – Helsinki Commission

**ICES WGMME:** International Council for the Exploration of the Seas Working Group on Marine Mammal Ecology

**ICES:** International Council for the Exploration of the Seas

IMO: International Maritime Organisation

**IOC**: International Oceanographic Commission

**MPA:** Marine Protected Area

**MSP:** Maritime Spatial Planning

**NATO:** North Atlantic Treaty Organisation

**ORE:** Offshore Renewable Energy

**OSPAR (Commission):** Oslo-Paris Commission

**OSPAR (Convention):** Convention for the Protection of the Marine Environment of the North-East Atlantic

**OWF:** Offshore Windfarm

PCB: Polychlorinated biphenyl

SEA: Strategic Environmental Assessment

TG NOISE: European Union Technical Group on Underwater Noise

UNECE: United Nations Economic Commission for Europe

**UNEP:** United Nations Environment Programme

VASAB: Visions and Strategies Around the Baltic Sea

VMS: Vessel Monitoring System

#### References

Agardy, T., Notarbartolo di Sciara & Christie, P. 2011. Mind the gap: Addressing the shortcomings of marine protected areas through large scale marine spatial planning, Marine Policy, 35, 2, 226-232.

Amaral, J.L., Frankel, A.S., Miller, J.H. et al. 2020. Bubble curtain effectiveness during impact pile driving for monopile installation at the Coastal Virginia Offshore Wind project, The Journal of the Acoustical Society of America, 148, 4, 2627-2628.

Arlidge, W.NS., Bull, J.W., Addison, P.F.E et al. 2018. A Global Mitigation Hierarchy for Nature Conservation, BioScience, 68, 5, 336-347.

ASCOBANS 1992. Agreement Text, https://www.ascobans.org/en/documents/agreement-text.

ASCOBANS 2009, Conservation Plan for Harbour Porpoises (Phocoena phocoena L.) in the North Sea 2009: <a href="https://www.ascobans.org/en/documents/action%20plans/North-Sea-Conservation-Plan">https://www.ascobans.org/en/documents/action%20plans/North-Sea-Conservation-Plan</a>

ASCOBANS 2016a. ASCOBANS Recovery Plan for Baltic Harbour Porpoises Jastarnia Plan (2016 Revision), 8th Meeting of the Parties to ASCOBANS, ASCOBANS Resolution 8.3 Helsinki, Finland, 30 August - 1 September 2016.

ASCOBANS 2016b. Resolution No. 5: Monitoring and Mitigation of Small Cetacean Bycatch. 8th Meeting of the Parties to ASCOBANS. Helsinki, Finland, 30 August–1 September 2016.

ASCOBANS 2020. Monitoring and Mitigation of Small Cetacean Bycatch, UNEP/ASCOBANS/Res.8.5(Rev.MOP9).

Bailey, H., Senior, B., Simmons, D. et al. 2010. Assessing underwater noise levels during pile-driving at an offshore windfarm and its potential effects on marine mammals, Marine Pollution Bulletin, 60, 7, 888-897.

Baltic Marine Environment Protection Commission (HELCOM) 2021. Baltic Sea Action Plan: 2021 Update, <u>https://helcom.fi/baltic-sea-action-plan/</u>.

Barlow, J. and Gisiner, R., 2006. Mitigating, monitoring and assessing the effects of anthropogenic sound on beaked whales, Journal of Cetacean Research and Management, 7, 3, 239-249.

Basran, C.J., Bertulli, C.G., Cecchetti, et al. 2019. First estimates of entanglement rate of humpback whales Megaptera novaeangliae observed in coastal Icelandic waters. Endangered species research, 38, 67-77.

Baulch, S. and Perry, C. 2014. Evaluating the impacts of marine debris on cetaceans. Marine Pollution Bulletin, 80: 210-221.

Bearzi, Giovanni, and Randall R. Reeves 2021. "Shifting baselines of cetacean conservation in Europe." ICES Journal of Marine Science 78, 7, 2337-2341.

Bellmann, M., May, A., Wendt, T. et al. 2020. Underwater noise during percussive pile driving: Influencing factors on piledriving noise and technical possibilities to comply with noise mitigation values: ERa Report: Experience report on pilingdriving noise with and without technical noise mitigation measures, itap, https://www.itap.de/en/publications/

Benhemma-Le Gall, A., Graham, I.M., Merchant, N.D. et al. 2021. Broad-Scale Responses of Harbor Porpoises to Pile-Driving and Vessel Activities During Offshore Windfarm Construction, Frontiers in Marine Science, 8, 664724.

Benjamins, S., Dale, A.C., Hastie, G. et al. 2015. Confusion reigns? A review of marine megafauna interactions with tidal-stream environments, Oceanography and Marine Biology: An Annual Review, 53, 1-54.

Benjamins, S., van Geel, N., Hastie, G., Elliott, J., and Wilson, B. (2016) Harbour porpoise distribution can vary at small spatiotemporal scales in energetic habitats. Deep Sea Research Part II: Topical Studies in Oceanography, 141, 191–202.

Bennett, N, J. 2019. Marine Social Science for the Peopled Seas, Coastal Management, 47, 2, 244-252.

Bennett, N.J., Roth, R., Klain, S.C. et al. 2017. Conservation social science: Understanding and integrating human dimensions to improve conservation, Biological Conservation, 205, 93-108.

Bernaldo de Quirós, Y., Fernandez, A., Baird, R.W., et al. 2019. Advances in research on the impacts of antisubmarine sonar on beaked whales. Proceedings of the Royal Society B, 286: 20182533.

Birdlife International 2022a. Are EU Member State's Maritime Spatial Plans Fit for Nature and Climate?: Technical Report – Approach and Main Findings, Birdife International, prepared by Walsh, C. & Loch, M, https://www.birdlife.org/news/2022/06/08/precious-space-maritime-spatial-plans-can-bring-back-nature-at-sea/.

BOEM 2014. Quieting Technologies for Reducing Noise During Seismic Surveying and Pile Driving Workshop Summary Report, OCS Report BOEM 2014-061.

Booth, C.G., 2020. Food for thought: Harbor porpoise foraging behavior and diet inform vulnerability to disturbance. Marine Mammal Science, 36, 1, 195-208.

Borges L. and Evans P.G.H. 1997. Spatial Distribution of the Harbour Porpoise and Fish Prey and their Associations in Southeast Shetland, Scotland. European Research on Cetaceans 10, 262-265.

Bossart, G.D. 2011, Marine Mammals as Sentinel Species for Oceans and Human Health, Veterinary Pathology, 48, 3, 676-690.

Brandt, M.J., Dragon, A.C., Diederichs, A., et al. (2018) Disturbance of harbour porpoises during construction of the first seven offshore wind farms in Germany. Marine Ecology Progress Series, 596, 213-232.

Branstetter, B.K., Bowman, V.F. et al. 2018. Effects of vibratory pile driver noise on echolocation and vigilance in bottlenose dolphins (*Tursiops truncatus*), The Journal of the Acoustical Society of America, 143, 1, 429-439.

Brownlow A., 2018. Update on UK and Irish Beaked Whale Unusual Mortality Event. 24th ASCOBANS Advisory Committee Meeting, Vilnius, 25 -27 September 2018.

Bundesministerium für Umwelt, Naturschutz und Reaktorsicherheit 2013. Konzept für den Schutz der Schweinswale vor Schallbelastungen bei der Errichtung von Offshore-Windparks in der deutschen Nordsee (Schallschutzkonzept), https://www.bsh.de/DE/THEMEN/Offshore/Meeresfachplanung/Flaechenentwicklungsplan/\_Anlagen/Downloads/FEP\_20 22\_2/Schallschutzkonzept\_BMU.html.

Burnham, R.E., Duffus, D.A. and Malcolm, C.D., 2021. Towards an enhanced management of recreational whale watching: The use of ecological and behavioural data to support evidence-based management actions. Biological Conservation, 255, 109009.

Burris, R.K. & Canter, L.W. 1997. Cumulative impacts are not properly addressed in environmental assessments, Environmental Impact Assessment Review, 17, 1, 5-18.

Campana, I., Crosti, R., Angeletti, D., et al. 2015. Cetacean response to summer maritime traffic in the Western Mediterranean Sea. Marine Environmental Research, 109, 1-8.

Carballeira Brana, C.B., Cerbule, K., Senff, P. et al. 2021. Towards Environmental Sustainability in Marine Finfish Aquaculture, Frontiers in Marine Science, 8, 666662.

Carlen, I., Nunny, L. & Simmonds, M. P. 2021. Out of Sight, Out of Mind: How Conservation Is Failing European Porpoises, Frontiers in Marine Science, 8, 617478.

Carlucci, R., Manea, E., Ricci, P. et al 2021. Managing multiple pressures for cetaceans' conservation with an Ecosystem-Based Marine Spatial Planning approach, Journal of Environmental Management, 287, 112240.

Carome, W., Slooten, E., Rayment, W., et al. 2022. A long-term shift in the summer distribution of Hector's dolphins is correlated with an increase in cruise ship tourism. Aquatic Conservation: Marine and Freshwater Ecosystems, 32, 10, 1660-1674.

Castellote, M. and Llorens, C. 2015. Review of the effects of offshore seismic surveys in cetaceans: are mass strandings a possibility? Advances in Experimental Medicine and Biology, 875, 133-143.

Chambault, P., Albertsen, C.M., Patterson, T.A., et al. 2018. Sea surface temperature predicts the movements of an Arctic cetacean: the bowhead whale. Scientific Reports, 8, 1, 1-12.

Chou, E., Southall, B.L., Robards, M. et al. 2021. International policy, recommendations, actions and mitigation efforts of anthropogenic underwater noise, Ocean & Coastal Management, 202, 1, 105427.

Clark, C.W. and Gagnon, G.C., 2006. Considering the temporal and spatial scales of noise exposures from seismic surveys on baleen whales. IWC/SC/58 E, 9.

Coates, D.A., Kapasakali, D-A., Vincx, M. et al. 2016. Short-term effects of fishery exclusion in offshore wind farms on macrofaunal communities in the Belgian part of the North Sea, Fisheries Research, 179, 131-138.

Conn, P.B. & Silber, G.K. 2013. Vessel speed restrictions reduce risk of collision-related mortality for North Atlantic right whales, Ecosphere, 4, 4, 1-16.

Constantine, R. and Bejder, L., 2007. Managing the whale-and dolphin-watching industry: time for a paradigm shift. In Marine wildlife and tourism management: Insights from the natural and social sciences (pp. 321-333). Wallingford UK: CAB International.

Convention on Biological Diversity 2022. Kunning-Montreal Global Biodiversity Framework, https://www.cbd.int/doc/decisions/cop-15/cop-15-dec-04-en.pdf

Convention on the Conservation of Migratory Species of Wild Animals (CMS) 1979. Convention Text, <u>https://www.cms.int/en</u>.

Cunen, C., Walløe, L., Konishi, K. and Hjort, N.L., 2021. Decline in body condition in the Antarctic minke whale (Balaenoptera bonaerensis) in the Southern Ocean during the 1990s. Polar Biology, 44, 2, 259-273.

Currie, J.J., McCordic, J.A., Olson, G.L., Machernis, A.F. and Stack, S.H., 2021. The impact of vessels on humpback whale behaviour: the benefit of added whale watching guidelines. Frontiers in Marine Science, 8, 601433.

Curtin, R. & Prellezo, R. 2010. Understanding marine ecosystem-based management: A literature review, Marine Policy, 34, 5, 821-830.

Dähne, M., Gilles, A., Lucke, K., Peschko, V., Adletr, S., Krügel, K., Sundermeyer, J., and Siebert, U. (2013) Effects of pile-driving on harbour porpoises (Phocoena phocoena) at the first offshore wind farm in Germany. Environmental Research Letters, 8: 025002.

Dannheim, J., Bergström, L., Birchenough, S.N.R., et al. 2020 Benthic effects of offshore renewables: identification of knowledge gaps and urgently needed research, ICES Journal of Marine Science, 77, 3, 1092-1108.

Das, K., Debacker, V., Pillet, S. and Bouquegneau, J.M., 2002. Heavy metals in marine mammals. In Toxicology of marine mammals (pp. 147-179). CRC Press.

Deppellegrin, D., Venier, C., Kyriazi, Z., et al. 2019. Exploring Multi-Use potentials in the Euro-Mediterranean sea space, Science of the Total Environment, 635, 612-629.

Di Clemente, J., Christiansen, F., Pirotta, E., et al. 2018. Effects of whale watching on the activity budgets of humpback whales, Megaptera novaeangliae (Borowski, 1781), on a feeding ground. Aquatic Conservation: Marine and Freshwater Ecosystems, 28, 4, 810-820.

Dolman, S. J., Evans, P. G. H., Ritter, F., Simmonds, M. P., and Swabe, J. (2021). Implications of new technical measures regulation for cetacean bycatch in European waters. Mar. Policy 124, 104320.

Dolman, S., Baulch S., Evans, P.G.H. et al. 2016. Towards an EU Action Plan on Cetacean Bycatch, Marine Policy, 72, 67-75.

Dolman, S.J. & Jasny, M.J. 2015. Evolution of Marine Noise Pollution Management, Aquatic Mammals, 41, 4,

Dolman, S.J., Pinn, E., Reid, R.J., et al. 2010. A note on the unprecedented strandings of 56 deep-diving whales along the UK and Irish coast. Marine Biodiversity Records, 3.

Duck, R.W. 2012. Marine Spatial Planning: Managing a Dynamic Environment, Journal of Environmental Policy and Planning, 14, 1, 67-79.

EA, 2019. Contaminants in Europe's Seas Moving Towards a Clean, Non-Toxic Marine Environment. European Environmental Agency Report No 25/2018.

Ehler, C. 2021. Two decades of progress in Marine Spatial Planning, Marine Policy, 132, 104134.

Ehler, C & Douvere, F. 2009. Marine spatial planning: a step-by-step approach toward ecosystem-based management. Intergovernmental Oceanographic Commission, <u>https://unesdoc.unesco.org/ark:/48223/pf0000186559</u>.

Eigaard, O.R., Bastardie, F., Hintzen, N.T. et al. 2017. The footprint of bottom trawling in European waters: distribution, intensity, and seabed integrity, ICES Journal of Marine Science, 74, 3, 847-865.

Elliot, B., Tarzia, M., Read, A.J. et al. 2023. Cetacean bycatch management in regional fisheries management organizations: Current progress, gaps, and looking ahead, Frontiers in Marine Science, 9, 1006894.

Elliott, B., 2020. A Review of Regional Fisheries Management Organization Efforts in Addressing Cetacean Bycatch: Report to the International Whaling Commission.

Erbe, C., Marley, S.A., Schoeman, R.P., et al. 2019. The effects of ship noise on marine mammals—a review. Frontiers in Marine Science, p.606.

ESPON 2020. Policy Brief: Maritime Spatial Planning and Land-sea Interactions, <u>https://www.espon.eu/brief-maritime-planning</u>.

European Commission 1992. Council Directive 92/43/EEC of 21 May 1992 on the conservation of natural habitats and of wild fauna and flora.

European Commission 2018. Cross-border Consultation on Maritime Spatial Plans: Final Technical Study, Written by the European MSP Platform under the Assistance Mechanism for the Implementation of Maritime Spatial Planning, https://maritime-spatial-planning.ec.europa.eu/sites/default/files/20190604\_cross-borderstudy\_published\_0.pdf.

European Commission 2020. Communication from the Commission to the European Parliament, the Council, the European Economic and Social Committee and the Committee of the Regions: EU Biodiversity Strategy for 2030 Bringing nature back into our lives.

European Commission 2020. Communication from the Commission to the European Parliament, the Council, the European Economic and Social Committee and the Committee of the Regions: An EU Strategy to harness the potential of offshore renewable energy for a climate neutral future.

European Commission 2023a. The EU blue economy report 2023, Publications Office of the European Union.

European Commission 2023b. Review on how to preserve space for the future uses of the seas: what methods can we apply to address the needs of future generations? Background Technical Study, Produced by the European MSP Platform under the Assistance Mechanism for the Implementation of Maritime Spatial Planning. https://cinea.ec.europa.eu/publications/review-how-preserve-space-future-uses-seas\_en

European Commission 2023c. EU Action Plan to Conserve Fisheries Resources and Protect Marine Ecosystems, https://ec.europa.eu/info/law/better-regulation/have-your-say/initiatives/12953-Action-plan-to-conserve-fisheriesresources-and-protect-marine-ecosystems\_en.

European Commission, European Climate, Infrastructure and Environment Executive Agency 2021. Guidelines for implementing an ecosystem-based approach in maritime spatial planning: including a method for the evaluation, monitoring and review of EBA in MSP, <u>https://data.europa.eu/doi/10.2926/84261</u>.

European Court of Auditors 2016. Special report no 03/2016: Combating eutrophication in the Baltic Sea: further and more effective action needed, <u>https://www.eca.europa.eu/en/publications?did=35757</u>.

European Parliament 2023. Amendments adopted by the European Parliament on 12 July 2023(1) on the proposal for a regulation of the European Parliament and of the Council on nature restoration (COM(2022)0304 – C9-0208/2022 – 2022/0195(COD)), https://www.europarl.europa.eu/doceo/document/TA-9-2023-0277\_EN.html#title1.

European Union 2001. Directive 2001/42/EC of the European Parliament and of the Council of 27 June 2001 on the assessment of the effects of certain plans and programmes on the environment.

European Union 2008. Directive 2008/56/EC of the European Parliament and of the Council of 17 June 2008 establishing a framework for community action in the field of marine environmental policy (Marine Strategy Framework Directive)

European Union 2014. Directive 2014/89/EU of the European parliament and of the council of 23 July 2014 establishing a framework for maritime spatial planning.

European Union 2014. Directive 2014/89/EU of the European Parliament and of the Council of 23 July 2014 establishing a framework for maritime spatial planning.

European Union 2019. Directive (EU) 2019/904 of the European Parliament and of the Council of 5 June 2019 on the reduction of the impact of certain plastic products on the environment.

Evans, P. & Waggitt J., 2020. Impacts of climate change on marine mammals, relevant to the coastal and marine environment around the UK, (MCCIP Science Review 2020). Marine Climate Change Impacts Partnership. <u>http://www.mccip.org.uk/media/2022/19\_marine\_mammals\_2020.pdf</u>.

Evans, P. 2018. Marine Protected Areas and marine spatial planning for the benefit of marine mammals, Journal of the Marine Biological Association of the United Kingdom, 98, 5, 973–976.

Evans, P.G.H. 2019. European whales, dolphins, and porpoises: marine mammal conservation in practice, Academic Press.

Evans, P.G.H. and Miller, L.A. 2004, Proceedings of the workshop on Active Sonar and Cetaceans, European Cetacean Society.

Evans, P.G.H., Baines, M.E., and Anderwald, P. 2011. Risk Assessment of Potential Conflicts between Shipping and Cetaceans in the ASCOBANS Region. ASCOBANS AC18/Doc. 6-04 (S).

Federal Public Service: Health, Food Chain Safety and Environment (Belgium) 2019. Annexe 2 Vision à long terme, objectifs et indicateurs, et choix stratégiques en matière d'aménagement PAEM 2020, <a href="https://www.health.belgium.be/nl/bijlage-2-mrp-langetermijnvisie-doelstellingen-en-indicatoren-en-ruimtelijke-beleidskeuzes">https://www.health.belgium.be/nl/bijlage-2-mrp-langetermijnvisie-doelstellingen-en-indicatoren-en-ruimtelijke-beleidskeuzes</a>.

Feingold, D. & Evans, P.G.H. 2014, Connectivity of Bottlenose Dolphins in Welsh Waters: North Wales Photo-Monitoring Report, Natural Resources Wales Research Report.

Fernandez-Betelu, O., Graham, I.M. & Thompson, P.M. 2022. Reef effect of offshore structures on the occurrence and foraging activity of harbour porpoises, Frontiers in Marine Science, 9, 980388.

Findlay, C.R., Rojano-Donate, L., Tougaard, J. et al. 2023. Small reductions in cargo vessel speed substantially reduce noise impacts to marine mammals, Science Advances, 9, 25, 2987.

Finneran, J.J. & Jenkins, A.K. 2012. Criteria and Thresholds for U.S. Navy Acoustic and Explosive Effects Analysis, Space and Naval Warfare Systems Center Pacific San Diego CA.

Fire, S.E., Flewelling, L.J., Stolen, M., et al. 2015. Brevetoxin associated mass mortality event of bottlenose dolphins and manatees along the east coast of Florida, USA. Marine Ecology Progress Series, 526, 241-251.

Fisher, C.R., Montagna, P.A. and Sutton, T.T., 2016. How did the Deepwater Horizon oil spill impact deep-sea ecosystems?. Oceanography, 29(3), pp.182-195.

Fontaine, M., Baird, S.J.E., Piry, S. et al. 2007. Rise of oceanographic barriers in continuous populations of a cetacean: the genetic structure of harbour porpoises in Old World waters, BMC Biology, 5, 30.

Ford, J.K., Ellis, G.M., Olesiuk, P.F. and Balcomb, K.C., 2010. Linking killer whale survival and prey abundance: food limitation in the oceans' apex predator? Biology letters, 6, 1, 139-142.

Fossi, M.C., Baini, M., Panti, C. and Baulch, S., 2018. Impacts of marine litter on cetaceans: a focus on plastic pollution, in Marine Mammal Ecotoxicology, 147-184.

Fraschetti, S., McOwen, Papa, L. et al. 2021 Where Is More Important Than How in Coastal and Marine Ecosystems Restoration, 8, 626843

Friess, B. & Gremaud-Colombier, M. 2021. Policy outlook: Recent evolutions of maritime spatial planning in the European Union, Marine Policy, 132, 103428.

Gabriele, C.M., Amundson, C.L., Neilson, J.L. et al. 2022. Sharp decline in humpback whale (Megaptera novaeangliae) survival and reproductive success in southeastern Alaska during and after the 2014–2016 Northeast Pacific marine heatwave, Mammalian Biology, 102, 1113-1131.

Gill, A.B. 2005. Offshore renewable energy: ecological implications of generating electricity in the coastal zone, Journal of Applied Ecology, 42, 4, 605-615.

Gilles, A., Scheidat, M. & Siebert, 2009. Seasonal distribution of harbour porpoises and possible interference of offshore wind farms in the German North Sea, Marine Ecology Progress Series, 383, 295-307.

Goertner, J.F. 1982, Prediction of Underwater Explosion Safe Ranges for Sea Mammals, Naval Surface Weapons Center, Silver Spring MD.

Götz, T. and Janik, V.M., 2013. Acoustic deterrent devices to prevent pinniped depredation: efficiency, conservation concerns and possible solutions. Marine Ecology Progress Series, 492, 285-302.

Gusatu, L.F., Menegon, S., Deppellegrin, D. et al. 2021. Spatial and temporal analysis of cumulative environmental effects of offshore wind farms in the North Sea basin, Nature: Scientific Reports, 11, 10125.

Halpern, B.S., Frazier, M., Potapenko, J. et al. 2015. Spatial and temporal changes in cumulative human impacts on the world's ocean, Nature Communications, 6, 7615.

Halpern, B.S., Walbridge, S., Selkoe, K.A. et al. 2008. A Global Map of Human Impact on Marine Ecosystems, Science, 319, 5865, 948-952.

Hammar, L., Molander, S., Palsson, J. et al. 2020. Cumulative impact assessment for ecosystem-based marine spatial planning, Science of the Total Environment, 734, 10, 139024.

Hanke, A.R., Kell, L.T. and Fortuna, C.M., 2020. A Review of Incidental Cetacean Bycatch Reporting in EU Waters. Collect. Vol. Sci. Pap. ICCAT, 77(4), pp.105-112.

Hassler, B., Gee, K., Gilek, M. et al. 2018. Collective action and agency in Baltic Sea marine spatial planning: Transnational policy coordination in the promotion of regional coherence, Marine Policy, 92, 138-147.

Haüssermann, V., Gutstein, C.S., Bedington, M., et al. 2017. Largest baleen whale mass mortality during strong El Niño event is likely related to harmful toxic algal bloom. PeerJ, 5, 3123.

Hazen, E.L., Scales, K.L., Maxwell, S.A. et al. 2018. A dynamic ocean management tool to reduce bycatch and support sustainable fisheries, Science Advances, 4, 5, 3001.

Heinis, F., de Jong, C.A.F. & von Benda-Beckmann, A.M. 2022. Framework for Assessing Ecological and Cumulative Effects 2021 (KEC 4.0) – Marine Mammals TNO 2021 R12503-UK.

HELCOM (2010) Hazardous substances in the Baltic Sea. An integrated thematic assessment of hazardous substances in the Baltic Sea. Baltic Sea Environment Proceedings No. 120B.

HELCOM (2018) State of the Baltic Sea. Second HELCOM holistic assessment 2011-2016. Baltic Sea Environment Proceedings No. 116B.

HELCOM & VASAB 2016. Guideline for the implementation of ecosystem-based approach in Maritime Spatial Planning (MSP) in the Baltic Sea area, <u>https://helcom.fi/helcom-at-work/groups/helcom-vasab-maritime-spatial-planning-working-group/</u>.

HELCOM-VASAB Maritime Spatial Planning Working Group 2021. Regional Maritime Spatial Planning Roadmap 2021-2030.

Hoarau, L., Dalleau, M., Delaspre, S., Barra, T. and Landes, A.E., 2020. Assessing and mitigating humpback whale (Megaptera novaeangliae) disturbance of whale-watching activities in Reunion Island. Tourism in Marine Environments, 153-4, 173-189.

Huang, L-F., Xu, X-M. Yang, L-L. et al. 2023. Underwater noise characteristics of offshore exploratory drilling and its impact on marine mammals, Frontiers in Marine Science, 10, 1097701.

ICES (2016a). Bycatch of small cetaceans and other marine animals—Review of national reports under Council Regulation (EC) No. 812/2004 and other information. ICES Special Request Advice Northeast Atlantic and Adjacent Seas Ecoregions.

ICES (2018). Report of the Working Group on Bycatch of Protected Species (WGBYC), 1-4 May 2018, Reykjavik. CES CM 2018/ACOM:25.

ICES (2022) ICES Stock Assessment Graphs and Data. Available to download from ICES website at <a href="https://www.ices.dk/data/assessment-tools">https://www.ices.dk/data/assessment-tools</a>

ICES. 2020. EU request on emergency measures to prevent bycatch of common dolphin (Delphinus delphis) and Baltic Proper harbour porpoise (Phocoena phocoena) in the Northeast Atlantic. In Report of the ICES Advisory Committee, 2020. ICES Advice 2020, sr.2020.04. https://10.17895/ices.advice.6023

IMO 2023. Addressing underwater noise from ships - draft revised guidelines and work plan agreed, Sub-Committee on Ship Design and Construction (SDC 9), 23-27 January 2023, https://www.imo.org/en/MediaCentre/MeetingSummaries/Pages/SDC-9.aspx.

nttps://www.imo.org/en/MediaCentre/MeetingSummaries/Pages/SDC-9.aspx.

International Council for the Exploration of the Seas Working Group on Marine Mammal Ecology (ICES WGMME) 2019. ICES Scientific Reports. 1:22., <u>https://ices-</u>

library.figshare.com/articles/report/Working\_Group\_on\_Marine\_Mammal\_Ecology\_WGMME\_/18618623.

International Oceanographic Commission & European Commission 2021. MSPglobal: international guide on marine/maritime spatial planning, https://unesdoc.unesco.org/ark:/48223/pf0000379196.

Jacob, C., van Bochove, J-W., Livingstone, S. et al. 2020. Marine biodiversity offsets: Pragmatic approaches toward better conservation outcomes, Conservation Letters, 13, 3, e12711.

Jepson, P.D., Deaville, R., Acevedo-Whitehouse, K., et al. 2013. What caused the UK's largest common dolphin (Delphinus delphis) mass stranding event?. PLoS One, 8, 4, 60953.

Jepson, P.D., Deaville, R., Barber, J.L., et al. 2016. PCB pollution continues to impact populations of orcas and other dolphins in European waters. Nature Scientific Reports 6, 18573.

Jones, F.C. 2016. Cumulative effects assessment: theoretical underpinnings and big problems, Environmental Reviews, 24, 2, 187-204.

Karkkainen, B.C. 2002. Toward a Smarter NEPA: Monitoring and Managing Government's Environmental Performance, 102, 4, 903-972.

Kastelein, R.A., Helder-Hoek, L., Van de Voorde, L., et al. 2017. Temporary hearing threshold shift in a harbor porpoise (Phocoena phocoena) after exposure to multiple airgun sounds. Journal of the Acoustical Society of America, 142: 2430.

Kavanagh, A.S., Nykänen, M., Hunt, W., Richardson, N. and Jessopp, M.J., 2019. Seismic surveys reduce cetacean sightings across a large marine ecosystem. Scientific reports, 9(1), 1-10.

Kebke, A., Samarra, F., & Derous, D. 2022. Climate change and cetacean health: impacts and future directions, Philosophical Transactions of the Royal Society B: Biological Sciences, 377, 1854.

Kershaw, J.L., Ramp, C.A., Sears, R. et al. 2021. Declining reproductive success in the Gulf of St. Lawrence's humpback whales (Megaptera novaeangliae) reflects ecosystem shifts on their feeding grounds. Global Change Biology, 27, 5, 1027-1041.

Ketten, D.R., 1995. Estimates of blast injury and acoustic trauma zones for marine mammals from underwater explosions. Sensory systems of aquatic mammals, 391-407.

Kidd, S., Shaw, D. & Janssen, H. 2019. Exploring land-sea interactions: Insights for shaping territorial space, Europa XXI, 36, 45-58.

Koschinski, S. & Lüdemann, K. 2020. Noise mitigation for the construction of increasingly large offshore wind turbines Technical options for complying with noise limits, commissioned by Bundesamt für Naturschutz.

Laist, D.W., Knowlton, A.R., Mead, J.G, Collet, A.S., and Podesta, M. 2001. Collisions between Ships and Whales. Marine Mammal Science, 17, 1, 35-75.

Lambert, E., Pierce, G.J., Hall, K. et al. 2014. Cetacean range and climate in the eastern North Atlantic: future predictions and implications for conservation, Global Change Biology, 20, 6, 1782-1793.

Leaper, R., MacLennan, E., Brownlow, A., et al. 2022. Estimates of humpback and minke whale entanglements in the Scottish static pot (creel) fishery. Endangered Species Research, 49, 217-232.

Lieberknecht, L. 2020 Ecosystem-Based Integrated Ocean Management: A Framework for Sustainable Ocean Economy Development, WWF Norway, <u>https://www.grida.no/publications/477</u>.

Loiseau, C., Thiault, L., Devillers, R. et al. 2012. Cumulative impact assessments highlight the benefits of integrating land-based management with marine spatial planning, Science of the Total Environment, 787, 147339.

Long, R.D., Charles, A. & Stephenson, R.L. 2015., Key principles of marine ecosystem-based management, Marine Policy, 57, 53-60.

Lusher, A.L., Hernandez-Milian, G., Berrow, S., Rogan, E. and O'Connor, J., 2018. Incidence of marine debris in cetaceans stranded and bycaught in Ireland: Recent findings and a review of historical knowledge. Environmental Pollution, 232, 467-476.

Lusseau, D., Kindt-Larsen, L. and van Beest, F.M., 2023. Emergent interactions in the management of multiple threats to the conservation of harbour porpoises. Science of the Total Environment, 855, 158936.

MacLeod, C.D., 2009. Global climate change, range changes and potential implications for the conservation of marine cetaceans: a review and synthesis. Endangered Species Research, 7, 2, pp.125-136.

Madsen, P.T., Wahlberg, M., Topugaard, J. et al. 2006. Wind turbine underwater noise and marine mammals: implications of current knowledge and data needs, Marine Ecology Progress Series, 309, 279-295.

Mann, J. and Teilmann, J. (2013) Environmental impact of wind energy. Environmental Research Letters, 8: 035001.

Marine Institute (2021). The Stock Book 2021: Annual Review of Fish Stocks in 2021 with Management Advice for 2022. Marine Institute, Galway, Ireland.

Marino, L., Gulland, F. and Parsons, E.C.M., 2012. Protecting wild dolphins and whales: current crises, strategies, and future projections. Journal of Marine Biology.

Martin, D.M. 2017. Ecological restoration should be redefined for the twenty-first century, Restoration Ecology, 25, 5, 668-673.

Maxwell, S.M., Hazen, E.L., Bograd, S.J. et al. 2013. Cumulative Impacts on marine predators, Nature Communications, 4, 2688.

Maxwell, S.M., Hazen, E.L., Lewison, R.L. et al. 2015. Dynamic ocean management: Defining and conceptualizing real-time management of the ocean, Marine Policy, 58, 42-50.

Mazzariol, S., Arbelo, M. Centelleghe, C. et al. 2018. Chapter 15 - Emerging Pathogens and Stress Syndromes of Cetaceans in European Waters: Cumulative Effects, Marine Mammal Ecotoxicology: Impacts of Multiple Stressors on Population Health, 401-428.

McGeady, R., McMahon, B.J. and Berrow, S., 2016, July. The effects of seismic surveying and environmental variables on deep diving odontocete stranding rates along Ireland's coast. In Proceedings of Meetings on Acoustics 4ENAL, 27, 1, 040006.

Meyer-Gutbrod, E.L., Greene, C.H., Davies, K.T.A, et al. 2021. Ocean Regime Shift is Driving Collapse of the North Atlantic Right Whale Population, Oceanography, 34, 3, 22-31.

Miller, P.J.O., Johnson, M.P., Biassoni, N., Quero, M., and Tyack, P.L. 2009. Using at-sea experiments to study the effects of airguns on the foraging behaviour of sperm whales in the Gulf of Mexico. Deep Sea Research Part I: Oceanographic Research Papers, 56, 7, 1168-1181.

Moodie, J.R. & Sielker, F. 2022. Transboundary Marine Spatial Planning in European Sea Basins: Experimenting with Collaborative Planning and Governance, Planning Practice & Research, 37, 3, 317-332.

Morkunas, J., Oppel S., Bruzas, M. et al. 2022. Seabird bycatch in a Baltic coastal gillnet fishery is orders of magnitude larger than official reports, Avian Conservation and Ecology, 1, 31, 1-12.

Murphy, S., Evans, P.G., Pinn, E. and Pierce, G.J., 2021. Conservation management of common dolphins: Lessons learned from the North-East Atlantic. Aquatic Conservation: Marine and Freshwater Ecosystems, 31, 137-166.

Murphy, S., Pinn, E. H., & Jepson, P. D. (2013). The short-beaked common dolphin (Delphinus delphis) in the Northeastern Atlantic: Distribution, ecology, management and conservation status. In R. N. Hughes, D. J. Hughes, & I.P. Smith (Eds.), Oceanography and marine biology: An annual review (Vol. 51) (pp. 193–280). Boca Raton, Florida: CRC Press.

Nachtsheim, D.A., Viquerat, S., Ramirez-Martinez et al. 2020. Small Cetacean in a Human High-Use Area: Trends in Harbor Porpoise Abundance in the North Sea Over Two Decades, Frontiers in Marine Science, 7, 606609.

National Academies of Sciences, Engineering and Medicine 2017. Approaches to Understanding the Cumulative Effects of Stresssors on Marine Mammals.

Naylor, W. & Parsons, E.C.M. 2018. An Online Survey of Public Knowledge, Attitudes, and Perceptions Toward Whales and Dolphins, and Their Conservation, Frontiers in Marine Science, 5, 00153.

New, L., Lusseau, D. and Harcourt, R., 2020. Dolphins and boats: when is a disturbance, disturbing? Frontiers in Marine Science, 7, 353.

New, L.F., Harwood, J., Thomas, L., et al. (2013) Modelling the biological significance of behavioural change in coastal bottlenose dolphins in response to disturbance. Functional Ecology, 27, 314-322.

Northridge, S., A. Kingston and L. Thomas (2018). Annual report on the implementation of Council Regulation (EC) No 812/2004 during 2017.

Northridge, S., Cargill, A., Coran, A., Mandleberg, L., Calderan, S., and Reid, R.J. (2010) Entanglement of minke whales in Scottish waters: an investigation into occurrence, causes and mitigation. Sea Mammal Research Unit, Final Report to Scottish Government CR/2007/49.

Nowacek, D., Thorne, L.H., Johnston, D.W., and Tyack, P.L. 2007. Responses of cetacean to anthropogenic noise. Mammal Review, 37(2): 81-115.

Nunny, L. & Simmonds, M.P. 2020. Climate Change and Cetaceans – an Update: Proceedings of the Annual Meeting of the International Whaling Commission Scientific Committee.

Olaya-Ponzone, L., Espada Ruiz, R., Paton Dominguez, D. et al. 2023. Sport fishing and vessel pressure on the endangered cetacean *Delphinus delphis*. Towards an international agreement of micro-sanctuary for its conservation, Journal of Environmental Management, 325, B, 116546.

Onoufriou, J., Brownlow, A., Moss, S., Hastie, G., and Thompson, D. 2019. Empirical determination of severe trauma in seals from collisions with tidal turbine blades. Journal of Applied Ecology, 56, 7, 1712-1724.

Orr, J.A., Vinebrooke, R.D., Jackson, M.C. et al. 2022. Towards a unified study of multiple stressors: divisions and common goals across research disciplines, Philosophical Transactions of the Royal Society B: Biological Sciences, 287, 1926.

OSPAR 2010a. Bergen Statement, Annex 49, Ref. M6.2.

OSPAR 2022. Pilot Assessment of Status and Trends of Persistent Chemicals in Marine Mammals, OSPAR Quality Status Report 2023, <u>https://oap.ospar.org/en/ospar-assessments/quality-status-reports/qsr-2023/indicator-assessments/pcb-marine-mammals-pilot/</u>.

OSPAR Commission 2010b. Quality Status Report 2010. OSPAR Commission, London.

OSPAR Commission 2014. Regional Action Plan for Prevention and Management of Marine Litter in the North-East Atlantic. OSPAR Commission, London.

OSPAR Commission 2017. Intermediate Assessment 2017. OSPAR Commission, London. https://oap.ospar.org/en/ospar-assessments/intermediate- assessment-2017

Ostend Declaration 2023. Ostend Declaration on the North Seas as Europe's Green Power Plant Delivering Cross-Border Projects and Anchoring the Renewable Offshore Industry in Europe, <u>https://windeurope.org/newsroom/press-</u> releases/eu-leaders-meet-in-ostend-to-agree-rapid-build-out-of-offshore-wind-in-the-north-seas/.

Pacheo, A.S., Sepulveda, M. & Corkson, P. 2021. Editorial: Whale-Watching Impacts: Science, Human Dimensions and Management, Frontiers in Marine Science, 8, 837352.

Parsons, E.C.M., 2017. Impacts of Navy sonar on whales and dolphins: Now beyond a smoking gun?. Frontiers in Marine Science, 295.

Parsons, E.C.M., Dolman S.J., Jasny, M.J. et al 2009. A critique of the UK's JNCC seismic survey guidelines for minimising acoustic disturbance to marine mammals: Best practise? Marine Pollution Bulletin, 58, 5, 643-651.

Parsons, E.C.M., Warburton, C.A., Woods-Ballard, A. et al. 2003. The value of conserving whales: the impacts of cetacean-related tourism on the economy of rural West Scotland, Aquatic Conservation, 13, 5, 397-415.

Pearson, H.C., Savoca, M.S., Costa, D.P. et al. 2022. Whales in the carbon cycle: can recovery remove carbon dioxide? Trends in Ecology and Evolution, 38, 3, 238-249.

Peel, D., Smith, J.N. & Childerhouse, S. 2018. Vessel Strike of Whales in Australia: The Challenges of Analysis of Historical Incident Data, Frontiers in Marine Science, 5, 00069.

Peltier, H., Beaufils, A., Cesarini, C. et al. 2019. Monitoring of Marine Mammal Strandings Along French Coasts Reveals the Importance of Ship Strikes on Large Cetaceans: A Challenge for the European Marine Strategy Framework Directive, Frontiers in Marine Science, 6, 00486.

Pérez-Jorge, S., Gomes, I., Hayes, K., et al. 2016. Effects of nature-based tourism and environmental drivers on the demography of a small dolphin population. Biological Conservation, 197, 200-208.

Picciulin, M., Armelloni, E., Falkner, R. et al. 2022. Characterization of the underwater noise produced by recreational and small fishing boats (<14 m) in the shallow-water of the Cres-Lošinj Natura 2000 SCI, Marine Pollution Bulletin, 183, 114050.

Pierce, G.J., Brownlow, A., Evans, P.G.H. et al 2022. Report of the ASCOBANS Resource Depletion Working Group, ASCOBANS/AC27/Doc.2.2. <u>https://www.ascobans.org/en/document/report-ascobans-resource-depletion-working-group-august-2022</u>.

Pierce, G.J., Weir, C.R., Gutierrez-Munoz, P. et al. 2020. Is Iberian harbour porpoise (Phocoena phocoena) threatened by interactions with fisheries? IWC Scientific Committee SC/68B/SM/04 Rev2, https://digital.csic.es/handle/10261/229202.

Pinarbasi, K., Galparsoro, I., Alloncle N. et al. 2020. Key issues for a transboundary and ecosystem-based maritime spatial planning in the Bay of Biscay, Marine Policy, 120, 104131.

Pirotta, E., Merchant, N.D., Thompson, P.M., et al. 2015. Quantifying the effect of boat disturbance on bottlenose dolphin foraging activity. Biological Conservation, 181: 82-89.

Platteeuw, M., Bakker, J., van den Bosch et al. 2017. A Framework for Assessing Ecological and Cumulative Effects (FAECE) of Offshore Wind Farms on Birds, Bats and Marine Mammals in the Southern North Sea, in Köppel, J. (ed) Wind Energy and Wildlife Interactions. Springer, Cham.

Pogoda, B., 2019. Current status of European oyster decline and restoration in Germany. Humanities, 8, 1, 9.

Quemmerais-Amice, F., Barrere, J., La Riviere, M. et al. 2020. A Methodology and Tool for Mapping the Risk of Cumulative Effects on Benthic Habitats, Frontiers in Marine Science, 7, 569205.

Reijnders, P.J.H. 1992. Harbour porpoises Phocoena phocoena in the North Sea: Numerical responses to changes in environmental conditions. Netherlands Journal of Aquatic Ecology, 28: 75-85.

Richardson, W.J., Greene, C.R., Malme, C.I., and Thomson, D.H. 1995. Marine Mammals and Noise. Academic Press, San Diego, CA.

Rijnsdorp, A.D. Bolam, S.G., Garcia, C., et al. 2018. Estimating sensitivity of seabed habitats to disturbance by bottom trawling based on the longevity of benthic fauna. Ecological Applications, 28(5): 1302–1312.

Ritter, F., & Panigada, S. 2019. Chapter 28 - Collisions of Vessels with Cetaceans—The Underestimated Threat, World Seas: An Environmental Evaluation (Second Edition) Volume III: Ecological Issues and Environmental Impacts, 531-547.

Ritter, F., de Soto, N.A. and Martín, V., 2019. Towards ship strike mitigation in the Canary Islands. In Proceedings of the Annual Scientific Committee Meeting Held in Nairobi in 2019.

Roach, M., Cohen, M., Forster, R. et al. 2018. The effects of temporary exclusion of activity due to wind farm construction on a lobster (*Homarus gammarus*) fishery suggests a potential management approach, ICES Journal of Marine Science, 75, 4, 1416-1426.

Ryan, C., Leaper, R., Evans, P.G.H., et al. (2016) Entanglement: an emerging threat to humpback whales in Scottish waters. Presented to the Scientific Committee Meeting of the International Whaling Commission, 2016, SC/66b/HIM/01.

Sanborn, T. & Jung, J. 2021. Intersecting Social Science and Conservation, Frontiers in Marine Science, 8, 676394.

Santora, J.A., Mantua, N.A, Schroeder, I.D. et al. 2020. Habitat compression and ecosystem shifts as potential links between marine heatwave and record whale entanglements, Nature Communications, 11, 536.

Schaffar AL, Madon BE, Garrigue CL, Constantine RO. (2009) Avoidance of whale watching boats by humpback whales in their main breeding ground in New Caledonia. IWC SC. 2009 Jun; 61: WW6.

Scheidat, M., Tougaard, J., Brasseur, S., Carstensen, J., van Polanen Petel, T., Teilmann, J.Schupp, M.F., Bocci, M., Deppellegrin, D. et al. 2019. Towards a Common Understanding of Ocean Multi-Use, Frontiers in Marine Science, 6, 165.

Scott, N.J. & Parsons, E.C.M. 2005. A survey of public opinion in south-west Scotland on cetacean conservation issues, Aquatic Conservation, 15, 3, 299-312.

Senigaglia, V., Christiansen, F., Bejder, L., et al. 2016. Meta-analyses of whale-watching impact studies: comparisons of cetacean responses to disturbance. Marine Ecology Progress Series, 542, 251-263.

Siebert, U., Wohlsein, P., Lehnert, K., Baumgärtner, W. 2007. Pathological Findings in Harbour Seals (Phoca vitulina): 1996–2005, Journal of Comparative Pathology, 137, 1, 47-58.

Silber, G.K., Adams, J.D. & Bettridge, S. 2012. Vessel operator response to a voluntary measure for reducing collisions with whales, Endangered Species Research, 17, 3, 245-254.

Singh, G.G., Lerner, J. Mach, M. et al. 2020. Scientific shortcomings in environmental impact statements internationally, People and Nature, 2, 2, 369-379.

Smith, G., LeTissier, M., O' Hagan, A.M. & Farrell, E.J. 2022. Policy Coherence for Climate Change Adaptation at the Land-Sea Interface in Ireland, Planning Practice & Research, 37, 2, 173-188.

Sonne, C., Siebert, U., Gonnsen, K. et al. 2020. Health effects from contaminant exposure in Baltic Sea birds and marine mammals: A review, Environment International, 139, 105725.

Sousa, A., Alves, F., Dinis, A. et al., 2019. How vulnerable are cetaceans to climate change? Developing and testing a new index, Ecological Indicators, 98, 9-18.

Southall, B.L., Finneran, J.J., Reichmuth, C., et al. 2019. Marine mammal noise exposure criteria: Updated scientific recommendations for residual hearing effects. Aquatic Mammals, 45, 2, 125-232.

Sparling, C., Smith, K., Benjamins, S., et al. 2015. Guidance to inform marine mammal site characterisation requirements at wave and tidal stream energy sites in Wales. NRW Evidence Report No. 82.

Stancheva, M. Stanchev, H., Zaucha, J. et al. 2022. Supporting multi-use of the sea with maritime spatial planning. The case of a multi-use opportunity development - Bulgaria, Black Sea, Marine Policy, 136, 104927.

Steinemann, A. 2000. Improving alternatives for environmental impact assessment, Environmental Impact Assessment Review, 21, 1, 3.21.

Steins, N.A., Veraart, J.A., Klostermann J.E.M. et al. 2021. Combining offshore wind farms, nature conservation and seafood: Lessons from a Dutch community of practice, Marine Policy, 126, 104371.

Stöber, U. & Thomsen, F. 2021. How could operational underwater sound from future offshore wind turbines impact marine life? Journal of the Acoustical Society of America, 149, 3. 1791-1795.

Stuiver, M., Soma, K., Koundouri, P. et al. 2016. The Governance of Multi-Use Platforms at Sea for Energy Production and Aquaculture: Challenges for Policy Makers in European Seas, Sustainability, 8, 4, 333.

Sutton, G., Folegot, T., Jessopp, M., Clorennec, D., 2013. Mapping the Spatiotemporal Distribution of Underwater Noise in Irish Waters (STRIVE 121). Environmental Protection Agency.

Sutton, M., Freake, M., Tepsich, P. and Moulins, A., 2021. Observations of marine species from whale watching vessels suggest that anthropogenic disturbance decreases biodiversity in the Pelagos Sanctuary. Bios, 92, 1, pp.8-16.

TG Noise 2022b. Setting of EU Threshold Values for continuous underwater sound: Recommendations from the Technical Group on Underwater Noise (TG Noise): MSFD Common Implementation Strategy Technical Group on Underwater Noise (TG NOISE) Deliverable 4 of the work programme of TG Noise 2022.

TG NOISE: 2022a. Setting of EU Threshold Values for impulsive underwater sound: Recommendations from the Technical Group on Underwater Noise (TG Noise) MSFD Common Implementation Strategy Technical Group on Underwater Noise (TG NOISE) Deliverable 2 of the work programme of TG Noise 2022.

Thompson, P.M., Brookes, K.L, Graham, I.M., et al. 2013. Short-term disturbance by a commercial two-dimensional seismic survey does not lead to long-term displacement of harbour porpoises. Proceedings of the Royal Society B, 280, 20132001.

Thomsen, F., 2009. Assessment of the environmental impact of underwater noise. OSPAR Commission. Biodiversity Series, 436.

Tougaard, J., Carstensen, J., Teilmann, J., Skov, H., and Rasmussen, P. 2009. Pile driving zone of responsiveness extends beyond 20 km for harbour porpoises (Phocoena phocoena (L.)). The Journal of the Acoustical Society of America, 126, 1, 11–14.

Tulloch, V.J.D., Plaganyi, E.E., Brown, C. et al. 2019. Future recovery of baleen whales is imperiled by climate change, Global Change Biology, 25, 4, 1263-1281.

Tyack, P.L., Johnson, M., Soto, N.A., Sturlese, A. and Madsen, P.T., 2006. Extreme diving of beaked whales. Journal of Experimental Biology, 209, 21, 4238-4253.

United Nations Economic Commission for Europe 2017. Convention on Environmental Impact Assessment in a Transboundary Context (Espoo Convention), Amended convention text, <u>https://unece.org/info/publications/pub/21618</u>.

Van Parijs, S.M., Clark, C.W., Sousa-Lima, R.S. et al. 2009. Management and research applications of real-time and archival passive acoustic sensors over varying temporal and spatial scales, Marine Ecology Progress Series, 395, 21-36.

van Weelden, C., Towers, J.R. & Bosker, T., 2021. Impacts of climate change on cetacean distribution, habitat and migration, Climate Change Ecology, 1, 100009.

Vanderlaan, A.S.M. and Taggart, C.T. 2007. Vessel collisions with whales: the probability of lethal injury based on vessel speed. Marine Mammal Science, 23, 1, 144-156.

Vanderlaan, A.S.M., Taggart, C.T., Serdynska, et al. 2008. Reducing the risk of lethal encounters: vessels and right whales in the Bay of Fundy and on the Scotian Shelf, Endangered Species Research, 4, 283-297.

Venier, C., Menegon, S., Possingham, H.P. et al. 2021. Multi-objective zoning for aquaculture and biodiversity, Science of the Total Environment, 785, 146997.

Waggitt, J., Dunn, H., Evans, P.G.H., Hiddink, J., Holmes, L., Keen, E., Murcott, B., Plano, M., Robins, P., Scott, B., Whitmore, J., and Veneruso, G. (2017) Regional-scale patterns in harbor porpoise occupancy of tidal stream environments. ICES Journal of Marine Science, 75, 2, 701-710.

Wann, S., Yang, X., Chen, X., et al. 2022. Emerging marine pollution from container ship accidents: Risk characteristics, response strategies, and regulation advancements, Journal of Cleaner Production, 376, 134266.

Ward, E.J., Holmes, E.E. and Balcomb, K.C., 2009. Quantifying the effects of prey abundance on killer whale reproduction. Journal of Applied Ecology, 46, 3, 632-640.

Walsh, C. 2021. Transcending Land-Sea Dichotomies through Strategic Spatial Planning. Regional Studies, 55, 5, 818-830.

Walsh, C., Sielker, F., Smith, G. & Crawford, J. (eds.) (2022) Planning for Sea Spaces I: Processes, Practices and Future Perspectives, Planning Practice and Research, 37, 2.

Waters, S. and Aggidis, G., 2016. A world first: Swansea Bay tidal lagoon in review. Renewable and Sustainable Energy Reviews, 56, 916-921.

Weilgart, L., 2017. Din of the deep: noise in the ocean and its impacts on cetaceans. In Marine Mammal Welfare, Springer, Cham, 111-124.

Williams, R., Cholewiak, D., Clark, C.W., et al. 2020. Chronic ocean noise and cetacean population models. J. Cetacean Res. Manage., 21, 1, 85-94.

Williamson, M.J., ten Doeschate, M.T.I, Deaville, R. et al. 2021. Cetaceans as sentinels for informing climate change policy in UK waters, Marine Policy, 131, 104634.

Winkler, C., Panigada, S., Murphy, S. and Ritter, F., 2019. Global Numbers of Ship Strikes: An Assessment of Collisions between Vessels and Cetaceans Using Available Data in the IWC Ship Strike Database. IWC B, 68.

Wisniewska, D.M., Johnson, M., Teilmann, J., et al. (2016) Ultra-high foraging rates of harbour porpoises make them vulnerable to anthropogenic disturbance. Current Biology, 26: 1-6.

Wright, A.J. & Cosentino, A. M. 2015. JNCC guidelines for minimising the risk of injury and disturbance to marine mammals from seismic surveys: We can do better, Marine Pollution Bulletin, 100, 1, 231-239.

Wright, A.J., Deak, T. and Parsons, E.C.M., 2011. Size matters: management of stress responses and chronic stress in beaked whales and other marine mammals may require larger exclusion zones. Marine Pollution Bulletin, 63, 1-4, 5-9.

Wright, A.J., Dolman S.J., Jasny, M. et al 2013. Myth and Momentum: A Critique of Environmental Impact Assessments, Journal of Environmental Protection, 2013, 4, 72-77.

Würsig, B. Greene, C.R. & Jefferson, T.A. 2000. Development of an air bubble curtain to reduce underwater noise of percussive piling, Marine Environmental Research, 49, 1, 79-93.

WWF 2021. Guidance Paper: Ecosystem-based Maritime Spatial Planning in Europe and how to assess it, https://www.wwf.eu/?4076441/Ecosystem-based-Maritime-Spatial-Planning-in-Europe-and-how-to-assess-it.

WWF 2022. Assessing the Balance between Nature and People in European Seas: Maritime Spatial Planning in the Baltic: Assessment Report. <u>https://www.wwf.eu/?6107066/Assessing-the-balance-between-nature-and-people-in-the-Baltic-Sea</u>.